

Estimation of Annual Effective Dose Due to Ingestion of Natural Radionuclides in Cattle in Tin Mining Area of Jos Plateau, Nigeria

Janet Ayobami Ademola

Department of Physics, University of Ibadan, Ibadan, Nigeria

Email: jaademola@yahoo.com, janet.ademola@mail.ui.edu.ng

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Abstract

Contamination of land could occur during the extraction and handling of materials containing high levels of naturally occurring radionuclide materials. These radionuclides find their ways into the food chain. The activity concentrations of natural radionuclides ^{40}K , ^{226}Ra , and ^{232}Th in heart, liver, kidney, lungs, spleen and beef of cattle slaughtered and consumed in a tin mining area of Jos Plateau, Nigeria were determined by gamma ray spectroscopy method. The activity concentration of ^{40}K is the highest in all the samples. The annual effective dose to man through the ingestion of the radionuclides in the organs was estimated. The mean annual effective doses calculated are 35.35 ± 13.84 , 57.89 ± 38.27 and $46.93 \pm 10.28 \mu\text{Sv}\cdot\text{y}^{-1}$ for heart, liver and kidney, respectively. Those of lungs, spleen and meat are 28.44 ± 15.70 , 48.34 ± 28.85 and $41.24 \pm 3.56 \mu\text{Sv}\cdot\text{y}^{-1}$, respectively. These are of the order of two magnitudes higher than those obtained for food in Abeokuta, Nigeria.

Keywords

Natural Radionuclides; Tin Mining; Contamination; Cattle; Ingestion; Effective Dose

1. Introduction

Water resources, agricultural areas, forest and the atmosphere are major biological systems determining the intake of radioactivity by man. Diet is the main source of radioactive elements, and thus of internal radiation dose [1]. A major source of energy for metabolic activities in living organisms is food. Ingestion of radionuclides through food intake accounts for substantial part of average radiation dose to various organs of the body and also represents one of the important pathways for long term health consideration. Radionuclides in soil may enter the food chain through direct deposition on leaves or through translocation to parts of plants that are edible to

humans and animals [2].

Radioactivity contaminated land can arise as a result of industrial processes, including past waste disposal or accident. Radionuclide could be present on land as a result of a number of past land uses. Contamination of land could occur during the extraction and handling of materials containing high levels of naturally occurring radionuclide materials (NORM) [3]. The intake or ingestion of radionuclide contaminated feed, e.g. grass by animals can concentrate radioactivity in tissues [4]. The uptake of radionuclides by animals is dependent on the animal species, mass, age, and growth rate of animal and the digestibility of the feed. Ingestion of soil during grazing can be an important contributor to intake of activity, particularly for radionuclides that are not taken up readily by grass [5]. Hence, contamination of meat is mainly the result of animal grazing, but contaminated drinking water might also be an important pathway [6]. People can therefore be exposed internally to radionuclides through eating the products of animals which have themselves eat contaminated substances.

The Jos area of Nigeria has been identified as a high background radiation [7] [8]. The waste products (tailing) from the mining of tin ore to obtain tin and columbite are discarded and dumped around the mining sites. Activity concentrations of natural radionuclides in tailings, soil and food in the Jos area had been reported [7]-[11]. The tailings have been found to contain high activity concentrations of ^{238}U and ^{232}Th [7] [12]. This could result in internal exposure of the populace through leached activity which may be directly ingested through drinking water or may indirectly enter the food chain by uptake through vegetation, fish, milk and meat. The objective of this work is to investigate the natural radioactivity levels (^{226}Ra , ^{40}K and ^{232}Th) in different organs of cattle grazed and slaughtered in the area for consumption and also evaluate the population radiation ingestion dose.

2. Material and Methods

2.1. Sample Collection and Preparation

Different organs of cattle, commonly eaten by the Nigerian populace were collected from an abattoir close to the tin mining area. The cattle slaughtered at the abattoir graze on grass and eat other feed made from combination of grains grown in the area. The organs include heart, liver, kidney, lungs, spleen and beef. All the organs were collected from the same cattle at a cost. The whole organs for each cattle were collected and labeled to avoid mix up. Because of the cost only five cattle were sampled. The samples were frozen before the period of analysis. The organs were oven dried at a temperature of 100°C until a constant weight was attained. The dried samples were crushed and grinded. They were weighed and packed in sample containers, sealed and kept for more than four weeks in order to for Ra and Th and their short-lived progenies to attain secular radioactive equilibrium.

2.2. Radioactivity Measurement

2.2.1. Determination of Natural Radionuclides

The radionuclide contents were determined using a lead-shielded $76\text{ mm} \times 76\text{ mm}$ NaI(Tl) detector crystal (Model No. 802 series, Canberra Inc.) coupled to a Canberra Series 10 plus Multichannel Analyzer (MCA) (Model No. 1104) through a preamplifier. The detector has a resolution of about 8% at 0.662 MeV of ^{137}Cs . This is capable of distinguishing the gamma ray energies considered during these measurements. The uranium-238 and thorium-232 activities were determined indirectly through the activities of their daughter products. The choice of the reference nuclides for their activity determination was made based on the fact that NaI(Tl) detector has a poor resolution, hence, the peaks of interest to be considered would be sufficiently discriminated and intense. Based on this consideration, therefore, the ^{226}Ra content of the samples was determined from the intensity of the 1760 keV gamma-ray peak from ^{214}Bi , the ^{232}Th content from the 2615 keV gamma-ray peak from ^{208}Tl and the ^{40}K content from 1460 keV gamma-ray peak following the decay of ^{40}K . These peaks are clean, reasonably strong with very low continuum. The energy calibration of the counting system was done as described in Ademola and Olatunji (2013), [13]. The sample container was placed on the detector and each sample was counted for 10 hours.

2.2.2. Determination of Activity Concentrations of ^{226}Ra , ^{232}Th and ^{40}K

The activity concentrations of the radionuclides were calculated by first determining the detection efficiency E_p of the system. The detection efficiency at the constant geometry of counting is defined as [14] [15];

$$E_p = \frac{A}{tCYm} \quad (1)$$

where A is the net area under the photopeak, C ($\text{Bq}\cdot\text{kg}^{-1}$) is the activity concentration of a reference sample of mass m (kg) counted for a time t (s) and Y is the gamma radiation yield. The efficiency was determined using a reference standard source of known activity concentrations, prepared from International Atomic Energy Agency (IAEA), Vienna, Austria and traceable to source Ref No. IAEA-312. The activity concentrations of each radionuclide in the samples were obtained by relating the detection efficiency to the net area under each photopeak.

3. Results and Discussion

The activity concentrations of the three radionuclides in the different organs followed no pattern. There are variations in the activity concentrations of ^{40}K , ^{226}Ra and ^{232}Th in the organs from one sample to the other. On the average the activity concentration of ^{40}K is higher in all the organs. The highest ^{40}K , ^{226}Ra and ^{232}Th contents were found in the liver of sample 2, heart of sample 5 and liver of sample 5, respectively. **Figures 1, 2 and 3** show the distribution of ^{40}K , ^{226}Ra and ^{232}Th , respectively in the different organs.

Radiation dose received due to the intake of food is calculated from the amount of radionuclide deposited on foodstuff, the activity concentration of particular radionuclide in food per unit deposition, the consumption rate of the food products and the dose per unit activity ingested [16]. Annual effective ingestion dose for an adult member of the public due to the intake of radionuclide through ingestion of food can be calculated based on the metabolic models developed by the International Commission of Radiological Protection [17];

$$H_{T,r} = \sum (U^i \times C_r^i) \times g_{T,r} \quad (2)$$

where, i denotes a food group, the coefficients U_i and C_r^i denote the consumption rate ($\text{kg}\cdot\text{y}^{-1}$) and activity concentration of the radionuclide r ($\text{Bq}\cdot\text{kg}^{-1}$), respectively, and $g_{T,r}$ is the dose conversion coefficient for ingestion of radionuclide r ($\text{Sv}\cdot\text{Bq}^{-1}$) in tissue T . For adult members of the public, the recommended dose conversion coefficient $g_{T,r}$ for ^{40}K , ^{226}Ra , and ^{232}Th , are 6.2×10^{-9} , 2.8×10^{-7} and 2.2×10^{-7} $\text{Sv}\cdot\text{Bq}^{-1}$, respectively [18]. Presently, no site specific consumption data exist in the study area and as a result the average per capita of beef consumption between 1961 and 2007 for Nigeria which is $0.98 \text{ kg}\cdot\text{y}^{-1}$ was used [19]. This value was adopted for the other organs since there is no such data for them. Equation (2) was used to calculate the effective dose due to the intake of beef and other organs.

The results of the effective dose due to intake are presented in column 5 of **Tables 1 to 5**. The effective dose

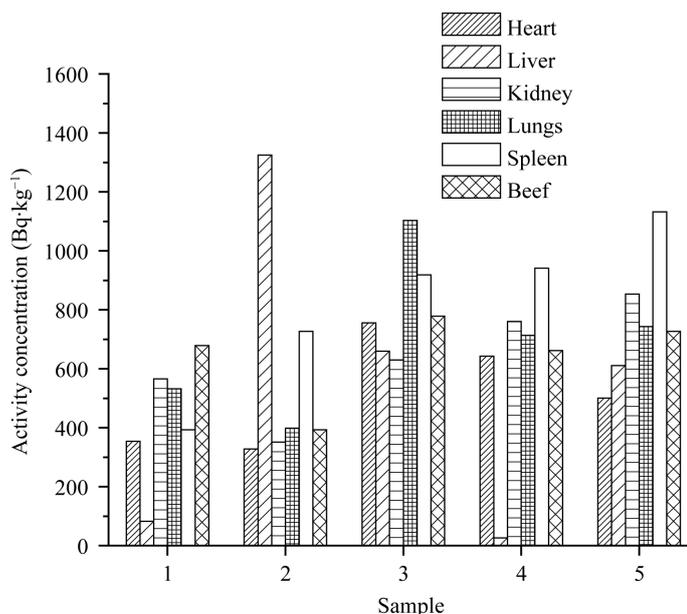


Figure 1. Activity concentration of ^{40}K in different organs of five cattle.

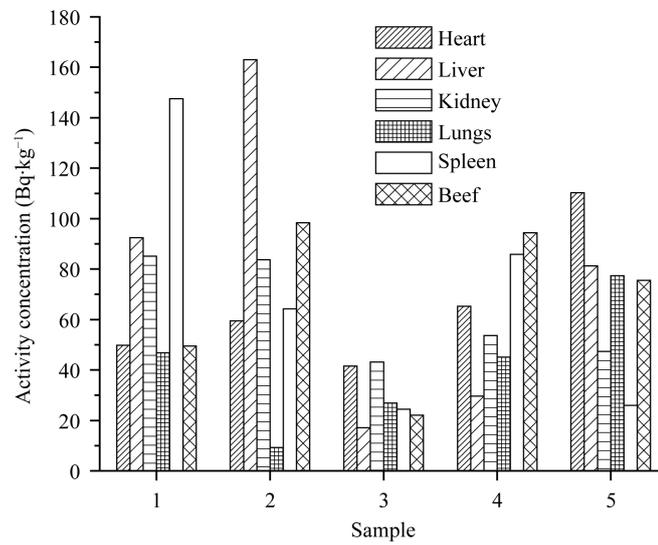


Figure 2. Activity concentration of ^{226}Ra in different organs of five cattle.

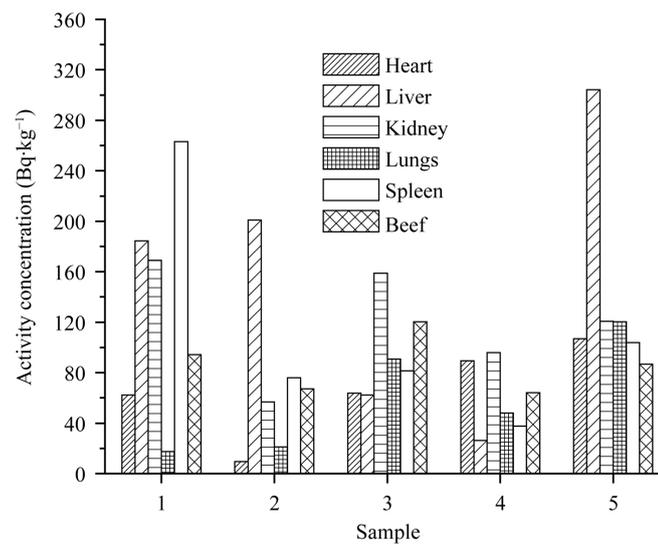


Figure 3. Activity concentration of ^{232}Th in different organs of five cattle.

Table 1. Activity concentrations of natural radionuclides and effective dose in different organs of sample 1.

Organ	^{40}K ($\text{Bq}\cdot\text{kg}^{-1}$)	^{226}Ra ($\text{Bq}\cdot\text{kg}^{-1}$)	^{232}Th ($\text{Bq}\cdot\text{kg}^{-1}$)	Effective dose ($\mu\text{Sv}\cdot\text{y}^{-1}$)
Heart	354.14	49.74	62.35	29.24
Liver	81.94	92.48	184.46	65.64
Kidney	565.8	85.15	169	63.24
Lungs	532.21	46.84	17.41	19.84
Spleen	392.38	147.55	263.13	99.6
Beef	678.67	49.49	94.29	38.03

Table 2. Activity concentrations of natural radionuclides and effective dose in different organs of sample 2.

Organ	^{40}K (Bq·kg ⁻¹)	^{226}Ra (Bq·kg ⁻¹)	^{232}Th (Bq·kg ⁻¹)	Effective dose (μSv·y ⁻¹)
Heart	327.89	59.45	9.58	20.37
Liver	1324.48	163.03	200.92	96.1
Kidney	351.44	83.65	56.71	37.32
Lungs	398.44	9.22	21.27	9.54
Spleen	726.86	64.28	75.93	38.43
Beef	393.12	98.39	67.17	43.87

Table 3. Activity concentrations of natural radionuclides and effective dose in different organs of sample 3.

Organ	^{40}K (Bq·kg ⁻¹)	^{226}Ra (Bq·kg ⁻¹)	^{232}Th (Bq·kg ⁻¹)	Effective dose (μSv·y ⁻¹)
Heart	755.96	41.49	63.81	29.74
Liver	660.08	17.13	62.4	22.16
Kidney	630.2	43.14	158.75	49.89
Lungs	1103.14	26.88	90.87	33.67
Spleen	918.12	24.42	81.35	29.89
Beef	778.13	22.06	120.33	36.72

Table 4. Activity concentrations of natural radionuclides and effective dose in different organs of sample 4.

Organ	^{40}K (Bq·kg ⁻¹)	^{226}Ra (Bq·kg ⁻¹)	^{232}Th (Bq·kg ⁻¹)	Effective dose (μSv·y ⁻¹)
Heart	642.43	65.24	89.3	41.06
Liver	25.86	29.65	26.36	13.98
Kidney	760.39	53.63	95.99	39.96
Lungs	713.21	45.12	48.04	27.07
Spleen	941.04	85.81	37.64	37.38
Beef	661.23	94.36	64.06	43.72

Table 5. Activity concentrations of natural radionuclides and effective dose in different organs of sample 5.

Organ	^{40}K (Bq·kg ⁻¹)	^{226}Ra (Bq·kg ⁻¹)	^{232}Th (Bq·kg ⁻¹)	Effective dose (μSv·y ⁻¹)
Heart	500.58	110.27	106.83	56.33
Liver	610.43	81.21	304.18	91.57
Kidney	853.69	47.37	120.74	44.22
Lungs	742.79	77.39	120.3	51.69
Spleen	1132.34	25.92	103.85	36.38
Beef	726.55	75.59	86.74	43.86

due to the ingestion of ^{40}K , ^{226}Ra and ^{232}Th through the consumption of the heart ranged between 20.37 and 56.33 μSv·y⁻¹ with a mean value of 35.35 ± 13.84 μSv·y⁻¹. That of liver varied from 13.89 to 96.10 μSv·y⁻¹ with a mean value of 57.89 ± 38.27 μSv·y⁻¹. Kidney, varied from 37.32 to 63.24 μSv·y⁻¹ with a mean value of 46.93 ± 10.28 μSv·y⁻¹. For lungs and spleen, the values ranged from 9.54 to 51.69 μSv·y⁻¹ and 29.89 to 99.60 μSv·y⁻¹ with mean values of 28.44 ± 15.70 μSv·y⁻¹ and 48.34 ± 28.85 μSv·y⁻¹. The beef varied between 36.72 to 43.87 μSv·y⁻¹ with a mean value of 41.24 ± 3.56 μSv·y⁻¹. The errors quoted are the standard deviations from the mean values which depict the variation of the effective dose within each organ. The standard deviation of the mean of the beef samples shows little variation in the effective dose. **Figure 4** shows a chart of the effective dose due to the ingestion of the radionuclide in the different organs. The results obtained in this study are of the order of 2 magnitudes higher than those obtained by Jibiri and Abiodun (2013), for food diet in Abeokuta, Southwestern

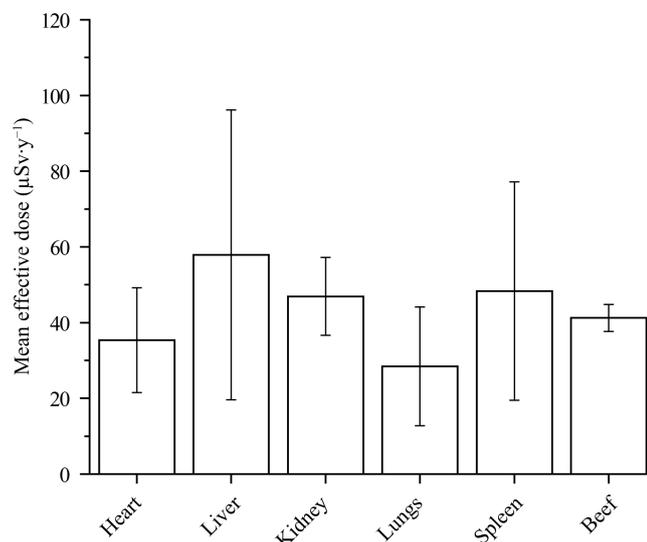


Figure 4. Mean annual effective dose due to the ingestion of ^{40}K , ^{226}Ra and ^{232}Th in different organs of cattle.

Nigeria [16].

4. Conclusion

The activity concentrations of natural radionuclides, ^{40}K , ^{226}Ra and ^{232}Th in different organs of cattle slaughtered and consumed in a high background radiation area had been measured using gamma ray spectroscopy method. On the average ^{40}K was found to be the highest in all the organs. The annual effective dose due the ingestion of the radionuclides in the organs was estimated. The mean annual effective varied from $28.44 \pm 15.70 \mu\text{Sv}\cdot\text{y}^{-1}$ (lungs) to $57.89 \pm 38.27 \mu\text{Sv}\cdot\text{y}^{-1}$ (liver).

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