

Calibration of SWAT2009 Using Crop Biomass, Evapotranspiration, and Deep Recharge: Calera Watershed in Zacatecas, Mexico Case Study

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ABSTRACT

Groundwater is the main source of water in the semi-arid Calera watershed, located in the State of Zacatecas, Mexico. Due to increasing population, rapid industrial growth, and increased irrigation to meet growing food demand, groundwater extraction in the Calera watershed are exceeding recharge rates. Therefore, development and evaluation of alternative water management strategies are needed for sustainable development of the region. The Soil and Water Assessment Tool (SWAT) model was selected for this purpose as it has been used to simulate a wide range of agricultural production, the extensive testing and application in diverse watersheds worldwide, and the potential for future linkage of this model to groundwater models. However, crucial flow data which are commonly used for calibrating hydrologic models are not available in this watershed. This paper describes a novel calibration methodology that uses biomass and water balance approach which has potential for calibration of hydrologic models in ungauged or data-scarce watersheds, which are prevalent in many parts of the world. Estimated long-term annual average actual evapotranspiration (AET), and deep aquifer recharge rates and plant biomass values based on the expert knowledge of researchers and managers in the watershed were used as targets for calibration. The model performance was assessed using the Nash-Sutcliffe efficiency coefficient (NSE), coefficient of determination (R^2), and percent bias (PBIAS, %) statistics. On average, the calibrated SWAT model yielded annual Nash-Sutcliffe efficiency coefficient values of 0.95, 0.99, and 0.85 for AET, recharge, and biomass, respectively. The coefficient of determination, values for AET, recharge, and biomass were 0.95, 0.94, and 0.99 respectively. The percent bias values of $\pm 2.21\%$, $\pm 0.18\%$, and $\pm 0.96\%$ for AET, recharge, and biomass, respectively, indicated that the model reproduced the calibration target values of the three water budget variables within an acceptable value of $\pm 10.0\%$. Therefore, it is concluded that the calibrated SWAT model can be used in evaluating alternative water management scenarios for the Calera watershed without further validation. Considering the relative ease in developing calibration data and excellent performance statistics, the calibration methodology proposed in this study may have the potential to be used for ungauged or data-scarce agricultural watersheds that are prevalent in many parts of the world.

Keywords: SWAT; Calera Watershed; Scenarios; Recharge; Evapotranspiration; Runoff; Erosion

1. Introduction

Underlying unconfined Calera aquifer is the primary source of water in the semi-arid Calera watershed, located in the most populated, economical and industrialized part of the state of Zacatecas, Mexico. Municipalities

within the Calera Watershed account for about 60% of the state population and about 80% of the Calera region population live in urban areas. Calera watershed provides 74% of the state's (GDP) (gross domestic product), and generate 17% of employment [1] with irrigated agriculture being the largest employer followed by commerce, mining, livestock, and industry [2]. The ma-

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major irrigated crops include red dry chili (*Capsicum annum*), corn (*Zea mays*), garlic (*Allium sativum*), alfalfa (*Medicago sativa*), onion (*Allium cepa*), dry beans (*Phaseolus vulgaris*), and livestock fodder.

Irrigated agriculture accounts for 80% of the total of the groundwater extraction. Urban and industrial uses, which are increasing at a rate of 0.595 Hm³/year [3], account for the remaining 20% of the total groundwater extraction, [4]. Further, low rainfall, low irrigation efficiency combined with poor land management practices are contributing to rapid declines in groundwater levels, which is indicated by cracks in the ground and collapse of areas where water has been over-exploited, [5]. Growth of urban centers and industries within the state in recent years have also contributed to overexploitation of the groundwater aquifer [6]. In the face of declining water supply and the demand for more food production to feed the growing population, the region also faces challenges posed by changing climate conditions caused by global warming [3,7,8].

Current and future water management in the Calera Watershed will determine the sustainability of irrigated agriculture in the region. It is therefore, important to develop a clear strategic plan to minimize overexploitation and to determine ways in which aquifer recharge can be enhanced. This task requires the cooperation of all stakeholders; farmers, municipal and industrial users, and water use policy and decision makers.

Hydrologic simulation models have proved to be useful tools to assess the effect of agronomic management practices on runoff, water quality, recharge, erosion and productivity [9,10]. These models also make it possible to evaluate cause-effect relationships without making changes in real systems [11]. One such model is the Soil and Water Assessment Tool (SWAT) [12]. SWAT has been successfully used to evaluate nonpoint-source water resource problems for a large variety of water quality applications globally [13,14]. SWAT contains a host of parameters, and Miller *et al.* [15] and Goodrich *et al.* [16], emphasize the importance of the process used for parameter estimation, because of the large degree of spatial variability in the topographic, soil, and land-cover characteristics within watersheds. Zhang *et al.* [17] highlighted that hydrologic models may contain parameters that cannot be measured directly due to measurement limitations and scaling issues; and that for practical applications in solving water resources problems these model parameters are calibrated to produce model predictions as close as possible to observed values.

Barlund *et al.* [18] applied the SWAT model to the Yläneenjoki River catchment draining to Lake Pyhäjärvi in the country of Finland. They found that SWAT can be calibrated for flow and sediment yield using catchment scale parameters. However, for nutrient concentrations SWAT parameters require some modifications in

order to describe correctly the reduction efficiency in local conditions. Rossi *et al.*, [19] calibrated SWAT model for the Leon River watershed in the central part of Texas, and indicated that the streamflow trends were simulated well ($0.65 < \text{ENS} \leq 0.75$ [good]) to very well ($\text{ENS} > 0.75$ [very good]) based on Moriasi *et al.* (2007) Nash-Sutcliffe efficiency (ENS) criteria. The average magnitude of streamflow simulations agreed well with observed values during the calibration phase ($\text{PBIAS} < \pm 10$ [very good]). However, the validation period agreement ($\text{PBIAS} \geq \pm 25$ [unsatisfactory]) was less than the rigor needed for the intended application. In another study, the SWAT model was calibrated and validated for streamflow for three watersheds in the Muscatatuck River Basin in southeast Indiana [20]. The streamflow predictions of the model were acceptable for monthly time step, however, the model performed somewhat poorly in predicting daily streamflow. The performance of the model in predicting groundwater table depth was not as good as for streamflow. However the model was able to predict the seasonal variation of groundwater table with correlation coefficients that varied from 0.46 to 0.88 in the calibration period and from 0.41 to 0.71 in the validation period. These results are similar to those found in Behera and Panda [21]; Bosch *et al.* [22]; Cheng *et al.* [23]; Du *et al.* [24]; Qi, and Grunwald [25]; Torres-Benites, *et al.* [10]; and Srinivasan *et al.*, [26].

The major water budget components in hydrological modeling are evapotranspiration (ET), soil moisture storage, and streamflow. Srinivasan *et al.*, [26] remarked that in large-scale watersheds, ET and soil water storage are not easy to extrapolate as there are not enough monitoring locations, but they suggested using crop yield or biomass that generally account for both ET and soil moisture. Recently, Torres-Benites [10] working in the watershed El Tejojote, Atlacomulco, State of Mexico showed that SWAT simulated grain yield of maize with reasonable precision (observed, 4302 kg/ha; simulated, 5104 kg/ha) and was more precise in simulating the annual water and sediment yield when calibrated against crop biomass/yield. With an uncalibrated SWAT model, Srinivasan, *et al.*, [26] showed that SWAT was able to predict corn and soybeans yields and ET satisfactorily over the long-term average for different sites of the Upper Mississippi River Basin. The PBIAS were less than $\pm 15\%$ and $\pm 10\%$ for crop yields and ET, respectively. On the other hand, they found that the PBIAS values can be larger than $\pm 20\%$, because much information on crop management is needed, in order to improve SWAT performance. Once the ET and crop yields are validated, SWAT model can provide the capacity to develop strategies for crop, soil and water management.

In Mexico, little effort has been made in using hydrologic models to develop an integrated strategy for quanti-

fication and control of soil erosion and water overexploitation. The long-term objective of this study is to develop a robust simulation framework to evaluate integrated policy, technological, and management strategies to enhance sustainability of the Calera Aquifer in the face of growing water demand and climate change. The SWAT model was selected because of the extensive validation and successful applications across a wide range of agroecological settings, the abundant crop, irrigation, and soil management features embedded within the SWAT databases, and the potential for linkage to other models within an integrated framework [10,23,25,26]. The specific objective of this paper is to document a detailed procedure to calibrate SWAT based on limited data on actual ET, recharge rate, crop biomass, and irrigation efficiency data for accurately predicting annual flow in the Calera watershed located in the State of Zacatecas, central Mexico.

2. Materials and Methods

2.1. Study Area

The Calera Aquifer is located between the geographic coordinates 22°38' to 23°15' latitude North, and 102°35' to 103°00' East in the central part of the Zacatecas State in north-central México. The 2056 km² watershed, representing 2.8% of the state area, **Figure 1** is a flat terrain that stretches from south to north without any obstruction, bordered on the east by the Zacatecas Mountains, the west by the Fresnillo and Los Cardos mountain chain, the north by the "Cerro del Algodón" and the south, by the Cerro de la Mesa.

2.2. Soil and Water Assessment Tool (SWAT) Overview

SWAT is a continuous-time, physically based, watershed-scale model developed to predict the impact of land management practices on water, sediment, and agricultural chemical yields in watersheds with varying soil, land use, and management conditions over time using information about weather, soils, topography, vegetation, ponds or reservoirs, groundwater, the stream network, and land management practices [27]. The model simulates flow, sediment, and nutrients at a watershed scale by dividing it into subbasins, which are further subdivided into homogeneous hydrologic response units (HRUs). These HRUs are the product of a distinct combination of soils and land use and there is routing from one HRU to another [12,13,26]. Hydrologic processes simulated by SWAT include surface runoff, infiltration, evapotranspiration, lateral flow, tile drainage, percolation and deep seepage, consumptive use through pumping, shallow aquifer contribution to streamflow (baseflow), and recharge by seepage from surface water bodies [27].

In this project, surface runoff and infiltration were estimated using the Soil Conservation Service (SCS) curve number procedure [28], and evapotranspiration (ET) was calculated using Penman-Monteith [29]. The ArcSWAT GIS (Geographic Information Systems) interface is used to facilitate hydrologic modeling tasks such as interposition of large layers of information at various spatial and temporal scales as stated above.

2.3. Modeling Approach

The modeling approach was divided into seven different phases: 1) collection and processing of basic information; 2) watershed delineation; 3) HRU analysis; 4) integration of input files; 5) management and reservoir data; 6) performance and calibration of the SWAT, and 7) preparing the document (**Figure 2**).

2.4. Conceptual Model

The Calera Aquifer underlying the Calera watershed is considered as an unconfined aquifer [30]. It was formed by alluvial and lacustrine deposits of clay, silt, sand, gravel, and gravel-sand conglomerates cemented with calcareous clay [31]. The basement depth is of approximately 500 m. The groundwater flows from south to north. The average depletion rate for the period of 1980 to 1994 was estimated between 0.4 and 1.15 m·y⁻¹ [2]; the recharge largely comes from rainfall and to a lesser extent of infiltration from anthropogenic irrigation practices. The Calera Aquifer is located at the north east of Zacatecas City, covering a surface of 1151 km², which corresponds to the 55% of the total geo-hydrological zone. It is elongated, with north-south orientation, 46 km long and 20 km wide, diminishing in width to the south and increasing it to the north. According to CONAGUA, equipotential lines of the static level converge at the center, and apparently, there is a way out towards the east to the Chupaderos Aquifer, while to the west, there is a zone of significant contribution from the Aguanaval to Calera Aquifer. Chupaderos and Aguanaval aquifers are contiguous (left and right, respectively) to Calera.

2.5. Weather Information

SWAT requires daily precipitation, maximum/minimum air temperature, solar radiation, wind speed and relative humidity. Values for all these parameters may be read from records of observed data or they may be generated [27]. In this study monthly precipitation and temperature data records were obtained from the CONAGUA weather stations of—Calera, and Fresnillo—for the period 1958-2010. The monthly observed precipitation and temperature data for 1958-2010 were used to generate the daily data externally. Then that generated daily data of precipitation and maximum and minimum temperature data

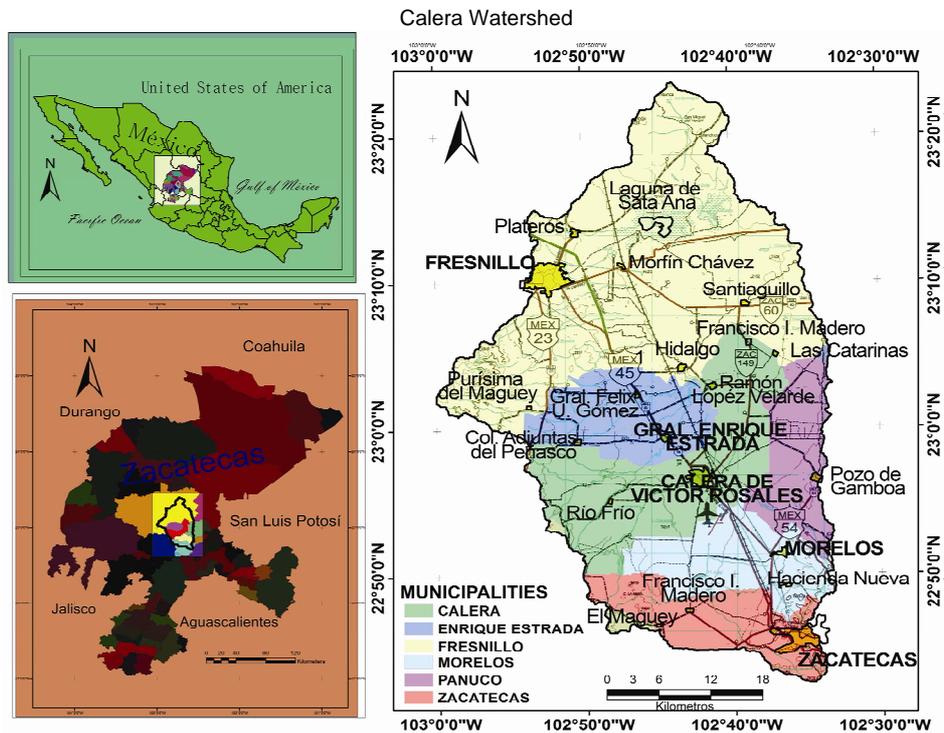


Figure 1. Calera watershed location and main urban developments.

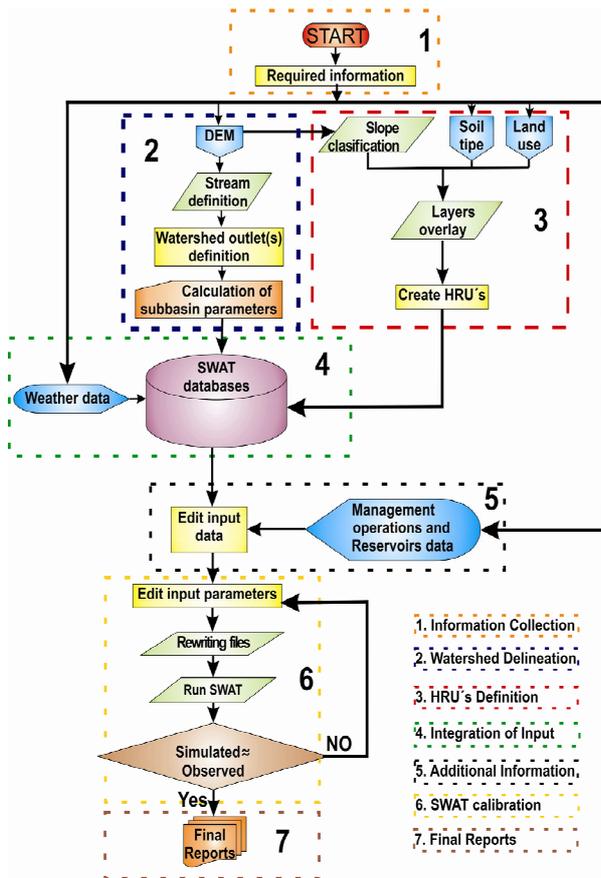


Figure 2. Methodology flowchart.

were supplied to the model using the ArcSWAT GIS interface. Wind speed, solar radiation, and air humidity were generated by WXGEN [32] weather generator in SWAT.

2.6. Soils

Soil texture and organic carbon data were obtained from the National Institute for Forestry, Agriculture and Livestock research (INIFAP, by the Spanish acronym). The Hydraulic Properties Calculator developed by Saxton and Rawls [33] was used to estimate soil available water, bulk density and saturated hydraulic conductivity data needed by SWAT based on the soils data from INIFAP. After this, GIS shape files with all the characteristics of the soil were created (Figure 3). This was done by associating generic (National Institute of Statistics and Geography, INEGI) data of predefined soil types in vector format with the soil texture, organic carbon, available water, bulk density and saturated hydraulic conductivity data using GIS tools.

2.7. Terrain and Slopes

A 30-meter digital elevation model (DEM) was obtained from INEGI and processed using the ArcSWAT interface, which allows the user to define the ranks of the slopes to be considered. In this study, four ranks were selected: 0% to 4%, 4% to 8%, 8% to 15%, and >15%. The predominant slope range was 0% to 4%, covering the valley which is

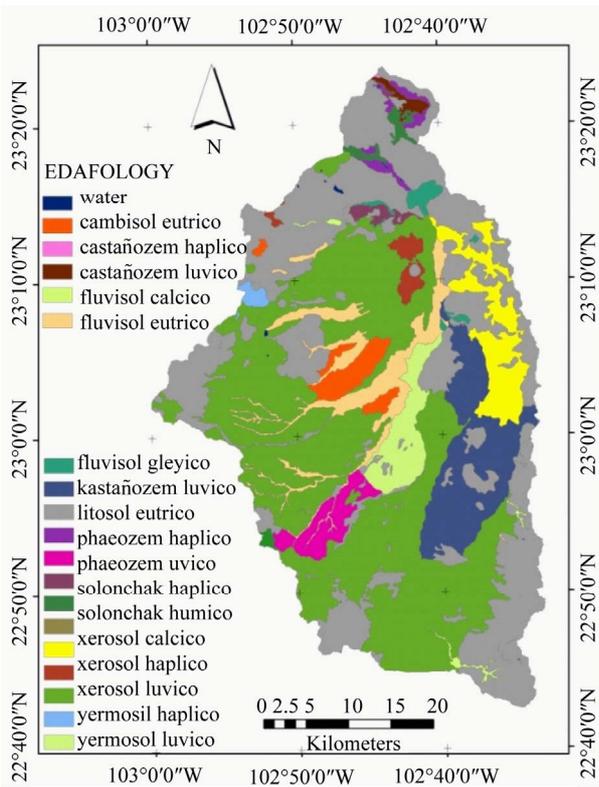


Figure 3. Map of soil types in the Calera watershed used in development of HRUs.

primarily cropland. The 4% to 8% slope range consists of primarily native grass lands; the 8% to 15% corresponds to the mountains and foothills; while slopes greater than 15% comprise highest areas of the mountains.

2.8. Land Use, Management, and Development of HRU's

A detailed land use map (**Figure 4**) for the study area was derived from a Landsat Thematic Mapper (TM) acquired on August 8, 2009. ERDAS Imagine (ERDAS Inc., 2010) was used for the classification of the image and geo-referencing the resulting land use map. An unsupervised classification technique was first employed to identify 150 statistical clusters and then regrouped into thematic classes such as corn, alfalfa, onion, garlic, red dry chili, dry beans, water, native vegetation, weeds/shrubs, urban/built-up area, and others. Ground truth data for assigning thematic classes to statistical clusters were collected during the 2009 growing season using a Global Positioning System (GPS) and overlaid on the clustered image to assign thematic classes. Finally, the land use map was geo-referenced to UTM coordinates to overlay with soil and slope layers for developing hydrologic response units (HRUs) in the ArcSWAT environment.

Agronomic management information related to the dates of planting, fertilization, irrigation and harvest for

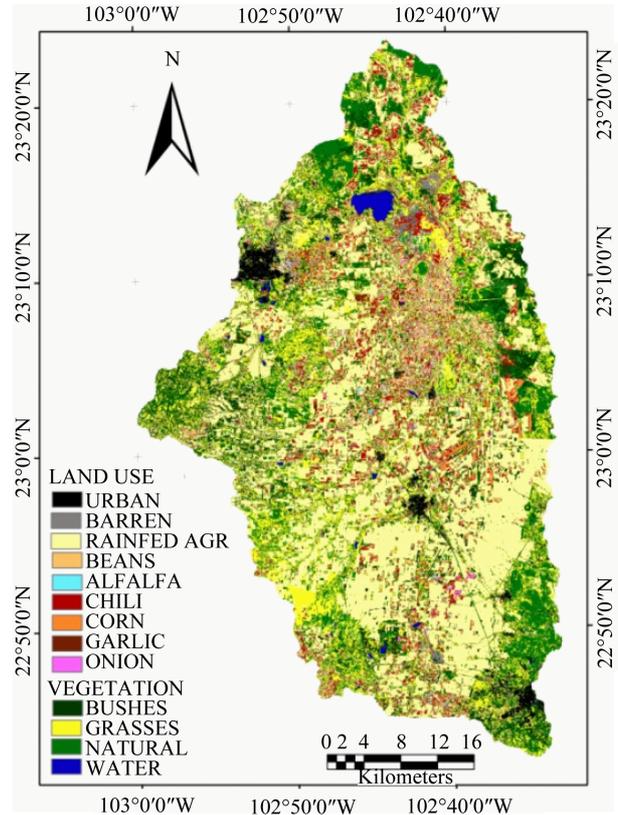


Figure 4. Map of the land use in the Calera watershed used in development of HRUs.

each thematic class was obtained from INIFAP and UAZ (Autonomous University of Zacatecas). The irrigation systems used include drip, sprinkler, gated furrow, and open ditch furrow. Because SWAT allows a maximum of 100 mm water per day, for the less efficient open ditch and gated pipe furrow irrigation systems in which applications greater than 100 mm of water per day were needed, 100 mm of water was applied on the recommended application date and the remaining water applied a day before or after.

2.9. Model Calibration

Due to the lack of available streamflow measurements for calibration and since the focus on irrigation and cropping system is the long-range goal of the project, evapotranspiration, plant biomass, and deep aquifer recharge information was used to calibrate the model and to ensure proper prediction of surface water processes and recharge to the groundwater. However, there was a lack of detailed watershed-scale measured data for the selected components. To achieve the calibration, we utilized a combination of measured and estimated values of average multi-year crop biomass, evapotranspiration (AET), and deep aquifer recharge to evaluate model performance on an annual time-step.

Measured biomass data (rainfed and the mean across irrigation systems) were obtained from historical measurements made by INIFAP and UAZ as part of their research experiments. For rainfed areas, the target AET was taken as 95% of precipitation based on recommendation by Villegas [34] for semi-arid zones. This target AET value was applied on rainfed crops, grass and natural vegetation HRUs. For irrigated crops, AET values were obtained from INIFAP, Mojarro and Bravo [35] and Avila [36]. Because the landscape characteristics of the southwestern USA are comparable to the Calera watershed, 3% of precipitation was assumed to recharge the deep aquifer [37]. In addition, 8% of irrigation water was assumed to recharge the deep aquifer based on information provided by CONAGUA. These expert knowledge based values were used to generate “observed” average annual deep aquifer recharge values for the rainfed and irrigated HRUs. Irrigated HRUs were further split as a function of the irrigation system based on the fraction of irrigated area under each system.

Based on recommendations in the SWAT2009 manual, the soil evaporation compensation coefficient (ESCO), the biomass-energy ratio (BIO_E) ($[\text{kg}\cdot\text{ha}^{-1}]/[\text{MJ}\cdot\text{m}^{-2}]$), and aquifer percolation coefficient (RCHRG_DP), were selected as the calibration parameters for AET, crop biomass, and deep recharge. The ESCO adjusts the depth distribution of soil evaporation to meet soil evaporative demand and varies between 0.01 and 1.0, both values inclusive. As the value of ESCO is reduced, the model is able to evaporate more water from deeper layers in the soil profile. The BIO_E is the amount of dry biomass produced per unit intercepted solar radiation in ambient CO_2 and varies between 10 and 90. The greater the BIO_E, the greater the potential increase in total plant biomass on a given day. The RCHRG_DP describes the fraction of percolation from the root zone that recharges the deep aquifer and varies between 0.0 (no percolation) and 1.0 (all the water percolating from the root zone reaches the deep aquifer).

Biomass, AET, and deep aquifer recharge calibration (1958-2010) at annual time step was accomplished by increasing or reducing the calibration parameter values within an acceptable range reported in the literature, one parameter at a time, until the calibration objective functions described below were met; default values were used for the rest of the parameters. Calibration was performed

for pasture/natural vegetation and each rainfed and irrigated crop HRUs. Irrigated crop HRUs were further split in proportion to the area under each irrigation system. For each HRU, the calibration parameters were adjusted until the simulated values were within 5% of the target AET and within 10% of the target deep aquifer recharge values and measured mean biomass.

One problem related to biomass calibration was the lack of crop parameters for red dry chili and garlic in the SWAT database [27]. Initially these parameters were estimated with values of the parameters perceived closest to the red dry chili and garlic but this resulted in significant underestimation of biomass for these crops. Similar problems had been observed by Crespo *et al.*, [38] in a study that demonstrated difficulties due to lack of the specific crop parameters to simulate the hydrologic performance of SWAT for forest micro-watersheds with mixed vegetation. Such uncertainty in the crop parameters can adversely impact the accuracy of the model performance as indicated by Moriasi and Starks [39]. Therefore, new values for Biomass, Harvest Index (HI), Leaf Area Index maximum (LAI_{max}), RUE (Radiation Use Efficiency), and Crop Cycle parameters for red dry chili and garlic were developed (**Table 1**) based on Amador *et al.*, [40]; Mojarro and Rincon [41], and the authors' expertise on the study area.

3. Model Performance Evaluation

Model performance, defined herein as the ability of SWAT to reproduce target average annual AET, biomass, and deep aquifer recharge during the calibration period is most often evaluated through both graphical comparisons and statistical tests. In this study, annual bar charts were used to identify model bias and differences in the magnitude of the target and simulated components. In addition, the percent bias (PBIAS, %) [42], and Nash-Sutcliffe efficiency [43], and coefficient of determination (R^2) statistics were used. The optimal value for PBIAS is zero; low values close to zero indicate a more accurate simulation model. Positive PBAIS values indicate underprediction and negative PBAIS values indicate overprediction by the model. PBIAS is calculated with Equation (1).

$$\text{PBIAS} = \frac{\sum_{i=1}^n (Y_i^{\text{obs}} - Y_i^{\text{sim}}) * (100)}{\sum_{i=1}^n (Y_i^{\text{obs}})} \quad (1)$$

Table 1. Crop parameters developed for use in the SWAT calibration.

Crop	Irrigation System	Crop cycle (days)	Yield (Kg/Ha)	Dry Biomass (Kg/Ha)	HI	LAI _{max}	RUE
Red Dry Chili	Furrow	150	2264	6860	0.34	2.85	2.7
	Drip	150	3529	7736	0.45	3.8	3.2
Garlic	Furrow	223	9411	2950	3.19	3.2	2
	Drip	223	9953	3120	3.19	4.2	2.9

where Y_i^{obs} is the “ i^{th} ” observed data point; Y_i^{sim} is the “ i^{th} ” simulated datapoint; and n the number of observations.

The NSE indicates how well the plot of observed versus simulated data fits the 1:1 line and also indicates the accuracy of the model in predicting observed values [43]. The NSE ranges between $-\infty$ and 1.0 with an NSE value of 1.0 being the optimal value. NSE values ≤ 0 indicate that the mean value of the observed time series is a better predictor than the simulated outputs. NSE is calculated as follows:

$$NSE = 1 - \frac{\sum_{i=1}^n (Y_i^{obs} - Y_i^{sim})^2}{\sum_{i=1}^n (Y_i^{obs} - Y^{mean})^2} \quad (2)$$

where NSE is normalized statistic that determines the relative magnitude of the residual variances between the observed and simulated values and Y^{mean} is the mean of the observed values. We used the **criteria** from Moriasi *et al.* [44] that we accepted the model calibration if the PBIAS is within 10% of the observed or target values and if the $NSE \geq 0.65$ on a monthly time step. But because we calibrated our model on the annual time-step, we set out threshold of $NSE \geq 0.75$ to show a satisfactory calibration. The R^2 according to Bravais-Person [45] is expressed as the squared ratio between the covariance and the multi-

plied standard deviations of the observed and simulated values. Values of R^2 range between 0 and 1 where value of 0 means no correlation, and a value of 1 means the observed dispersion is equal to the simulated dispersion. R^2 is calculated as:

$$R^2 = \left[\frac{\sum_{i=1}^n (Y_i^{obs} - Y^{mean})(Y_i^{sim} - Y^{mean})}{\sqrt{\sum_{i=1}^n (Y_i^{obs} - Y^{mean})^2} \sqrt{\sum_{i=1}^n (Y_i^{sim} - Y^{mean})^2}} \right]^2 \quad (3)$$

These composite statistics in addition to the measured and simulated mean and standard deviation were computed using the calibrated values for the pasture/natural vegetation and each rainfed and irrigated crops.

4. Results and Discussion

Figures 5-7 illustrate AET, deep aquifer recharge, and biomass simulation performance, respectively, on an annual time step for the Calera watershed during the calibration period for each individual landuse and irrigation type. The differences are expressed as a percent of the observed or target values (Difference = ((Obs - Sim)/(Obs))*100. The corresponding calibration statistics are given in **Table 3**. The general annual graphical and statistical calibration criteria indicate

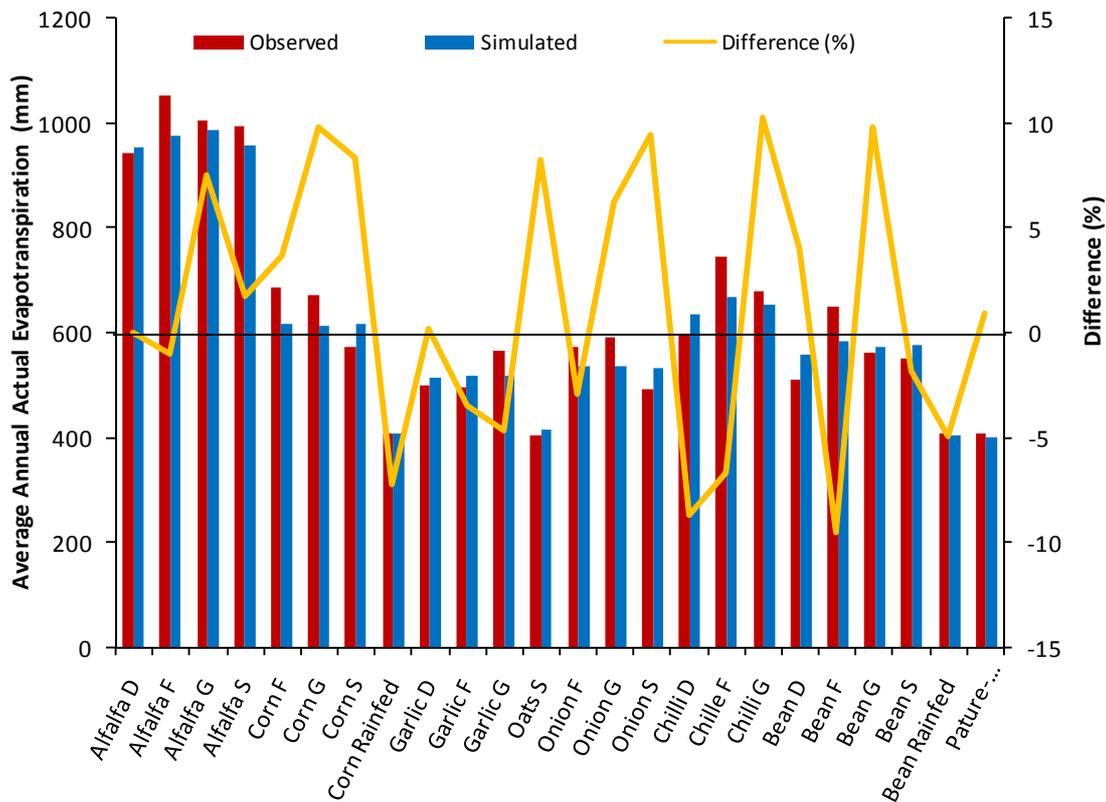


Figure 5. Target and simulated average annual AET values by crop and irrigation type (Drip = D; Sprinkler = S; Furrow Gated Pipe = G; Furrow Opened Ditch = F).

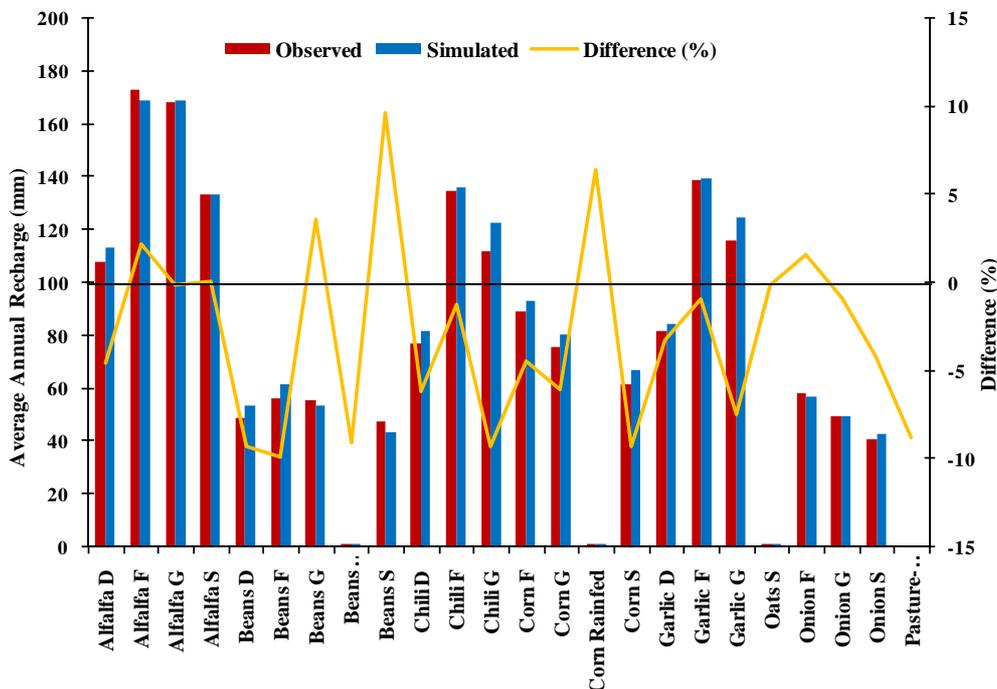


Figure 6. Target and simulated average annual deep aquifer recharge by crop and irrigation type (Drip = D; Sprinkler = S; Furrow Gated Pipe = G; Furrow Opened Ditch = F).

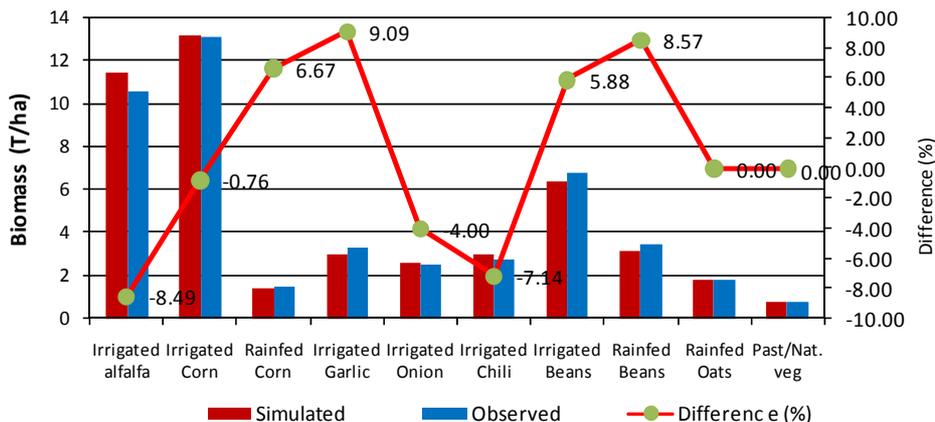


Figure 7. Target and simulated average annual biomass values by crop and presence or absence of irrigation.

that the model predicted AET, deep aquifer recharge, and biomass well according to Moriasi *et al.* [44] criteria.

Simulated AET values for the irrigated and rainfed crops (Figure 5) matched target values very well (-9.5% to +9.9%). With the exception of pasture/natural vegetation, rainfed HRUs were within 1.0% of the target average annual AET, which is equal to 95% of the precipitation. These results are similar to findings of Arnold and Allen [46] (1996) whose simulated AET values for the rainfed crops were within 1.0% of the target long-term average annual AET for three Illinois watersheds. Although the predicted AET for pasture/natural vegetation HRU was not within the 1.0% as in case of other land uses, it was still within the range of ±5.0% of the target

AET value of 95% of precipitation. Simulated AET for irrigated crops was within 10% of target ET values.

The simulated deep aquifer recharge rates (Table 2, and Figure 6) were comparable to those reported by Scanlon, *et al.*, [37] based on 140 study areas ranging between 40 and 374,000 km² in semiarid and arid regions of the world. They found that deep aquifer recharge values ranged from 0.1 to 35 mm/year representing 1% - 5% of long-term average annual precipitation. In irrigated areas Scanlon, *et al.* [37] found that deep aquifer recharge varied from 10 to 485 mm/year, 1% - 25% of irrigation plus precipitation. In desert sites, Gee *et al.*, [47] and Wang *et al.*, [48] found that deep aquifer recharge in non-vegetated areas were up to 87 mm-year⁻¹ with no

Table 2. Mean, standard deviation (STDEV), PBIAS, NSE, and R² for evapotranspiration (AET), aquifer recharge, and biomass in the calibration period (1952-2010).

Variable	Mean		Stdev		PBIAS (%)	NSE	R ²
	Target	Sim	Target	Sim			
Evapotranspiration (mm)	628.48	614.59	195.2	180.9	2.21	0.95	0.95
Aquifer Recharge (mm)	82.58	82.43	54.79	53.53	0.18	0.99	0.94
Biomass (Mg/Ha)	4.16	4.16	4.66	4.66	-0.96	0.85	0.99

recharge in vegetated areas. More recently in India, Scanlon, *et al.* [49] found that minimum deep aquifer recharge rates in irrigated agricultural sites were 50 to 120 mm/year.

Figure 6 illustrates the comparison between measured and simulated annual average deep aquifer recharge and the corresponding percent differences for each land use. For irrigated systems, results show higher recharge values for less efficient irrigation systems such as furrow open ditch and furrow gated compared with drip and sprinkler. However, there is savings in groundwater extractions when more efficient irrigation systems are used. It is important that water managers take these factors into consideration when they make decisions and policy regarding water sustainability.

For individual crops under different irrigation systems, percent differences between target and simulated biomass ranged from -8.49% to +9.9% (**Figure 7**) with mean and standard deviation biomass values of 4.16 and 4.66 Mt/ha (**Table 2**), respectively. Overall, SWAT simulated biomass very well, within $\pm 1.0\%$ (**Table 2**). Previous studies using SWAT also showed that the model adequately simulated crop biomass [10, 21, 26].

Values for the calibration parameters obtained through the calibration process are given in **Table 3**. The deep aquifer recharge and biomass parameters varied considerably with land use and irrigation system, with RCHRG_DP ranging from 0 to 0.25 and BIO_E ranging from 6 to 90 while ESCO was same for all land use and irrigation systems. All other parameters were held at the default values.

5. Conclusion and Recommendations

This study presented a detailed procedure to calibrate the SWAT model for predicted annual flow in the Calera watershed based on crop biomass, AET, and deep aquifer recharge. Based on PBIAS, the model simulated all the three variables reasonably well; simulated AET values were within 3% of the measured values and biomass and deep aquifer recharge values were within 1% of the observed values. Also calibration parameters for red dry chili and garlic, which are currently not available in the SWAT crop parameter datasets, were developed in this study. Incorporation of these parameter values into the

Table 3. Values of the calibrated parameters in SWAT.

Land Use/Irrigation System	ESCO	RCHRG_DP	BIO_E
Alfalfa/Drip irrigation	0.01	0.23	6
Alfalfa/Furrow opened ditch	0.01	0.13	6
Alfalfa/Furrow gated pipe	0.01	0.15	6
Alfalfa/Sprinkle	0.01	0.2	6
Corn/Furrow opened ditch	0.01	0.2	90
Corn/Furrow opened ditch	0.01	0.17	90
Corn/Sprinkle	0.01	0.2	90
Corn/Rainfed	0.01	0	8
Garlic/Drip irrigation	0.01	0.15	90
Garlic/Furrow opened ditch	0.01	0.13	90
Garlic/Furrow gated pipe	0.01	0.15	90
Onion/Furrow opened ditch	0.01	0.17	60
Onion/Furrow gated pipe	0.01	0.2	60
Onion/Sprinkle	0.01	0.22	60
Dry pepper/Drip irrigation	0.01	0.18	90
Dry pepper/Furrow opened ditch	0.01	0.12	90
Dry pepper/Furrow gated pipe	0.01	0.13	90
Dry beans/Furrow opened ditch	0.01	0.13	10
Dry beans/Furrow gated pipe	0.01	0.18	10
Dry beans/Sprinkle	0.01	0.2	10
Dry beans/Rainfed	0.01	0	19
Oat/Rainfed	0.01	0	15
Natural vegetation (arid zone)	0.01	0	18
Shrubs and low vegetation	0.01	0	10
Grasses	0.01	0	10

SWAT input datasets will ensure that future SWAT users in areas with these crops are grown will have the information they need. Based on the calibration results, we feel confident that the ET, deep aquifer recharge, and biomass parameter values obtained for each land use are generally representative of the hydrologic processes in the Calera watershed. The calibrated SWAT model will be used in the next phase to evaluate various irrigation and land use management systems to determine the best scenarios to minimize groundwater depletion rates. Also the acceptable model results for the Calera watershed show that the proposed calibrations approach has poten-

tial for use in many ungauged and data scarce watersheds worldwide.

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