Indoor Air Quality in Central Appalachia Homes Impacted by Wood and Coal Use

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ABSTRACT

Though the high prevalence of biomass fuel use in the developing world is widely known, the use of burning biomass for cooking and heating in the developed world is under-recognized. Combustion materials including coal and wood are also used for heating in some areas of the United States. We conducted a pilot study to assess the feasibility of conducting indoor environmental monitoring in rural Appalachia. We sought to explore the type of biomass being used for home heating and its impact upon indoor air quality in non-heating and heating seasons. Residential indoor air monitoring for particulate matter (PM) and nitrogen dioxide (NO₂) was conducted in Lee County, Virginia. Homes had evidence of poor indoor air quality with high concentrations of indoor PM and a large burden of cigarette smoking. Further characterization of indoor combustion material use in this region to determine the health impacts associated with such exposures is warranted.

Keywords: Indoor Air Pollution; Biomass; Rural Environment

1. Introduction

Over half of the world's population relies on open burning of biomass fuels for cooking and heating [1]. The World Health Organization estimates that emissions from cook stoves are one of the top five threats to public health in poor, developing countries [2]. Burning biomass significantly alters the indoor environment by increasing levels of pollutants, including particulate matter (PM), which are linked to poor health outcomes [2-6]. Although biomass fuels are generally perceived to be a public health problem in developing countries, indoor combustion materials, including coal and wood, are also used as primary or secondary heating fuels in many communities in the United States.

In some areas of Central Appalachia, it is estimated that up to 30% of homes use wood and/or coal for primary heating [7]. Cultural differences and housing characteristics may alter the impact of fuel use on the concentration of indoor pollutants, as pollutant burden is dependent on episodic characteristics of fuel burning and microenvironments around the fuel, including ventilation [8-10]. Consequently, the effect of coal and wood use on indoor pollutant concentrations in developed nations may differ from that observed in developing countries.

The impact of wood and/or coal burning on the indoor environment in Appalachia is not well described, and we are unaware of any recent studies that have conducted direct indoor air monitoring in residential settings in this region. To address this gap in knowledge, we designed a pilot study to assess the feasibility of conducting indoor environmental monitoring in rural Appalachia, and to explore the type of fuel being used for home heating and its impact upon indoor air quality in non-heating and heating seasons.

2. Methods

2.1. Study Design

Recruitment began in July of 2011 and ended in December 2011. Homes in Lee County, Virginia were recruited in partnership with the Appalachia Service Project (ASP), a faith-based, non-profit Christian organization that provides free home repair to needy families in the Central



Appalachia region. Families receiving services from ASP were notified of the study by ASP staff and contacted by study personnel to arrange for sampling. Owner-occupied homes with adequate electricity and an English-speaking homeowner able to provide informed consent were eligible. The study was reviewed and approved by the Johns Hopkins Institutional Review Board. Written informed consent was obtained from home owners.

2.2. Environmental Sampling

Assessments were conducted during a warm weather season (Visit 1, August 2011) and cold weather season after indoor home heating had commenced (Visit 2, December 2011). Each visit consisted of home inspection and indoor air monitoring including sampling for PM_{2.5} and coarse PM (PM_{2.5-10}; calculated by subtracting PM_{2.5} from PM₁₀), air nicotine, and nitrogen dioxide (NO₂) in the main living area of the home. Air sampling was conducted in the main living area of the home over a sampling period of four days. Continuous air sampling for particulate matter was conducted in the active mode, using PM₁₀ and PM_{2.5} 4 L/min impactors (MSP Corp, St. Paul, MN) loaded with 37 mm, 2.0 um pore size, Teflon polytetrafluoroethylene membrane filters with polypropylene support rings (Pall Corporation, Ann Arbor, MI) and BGI 400S sampling pumps (BGI, Inc., Waltham, MA). Four-day integrated NO₂ samples were collected using a passive sampler (Ogawa & Co., Pompano Beach, FL) loaded with filters coated with triethanolamine (TEA). In the presence of a color reagent, NO₂ and TEA form a highly colored azo dye that is measured spectrophotometrically at 540 nm. The median limit of detection (LOD) was 2 ppb. All analytical batches included 10% field and laboratory blanks, duplicates, and spiked samples.

Airborne nicotine (second hand smoke; SHS) was sampled using a passive air sampler consisting of a sodium bisulfate-treated filter contained in a 37-mm polystyrene cassette covered with a polycarbonate filter diffusion screen [11]. Nicotine content was analyzed using gas chromatography with a GC/MS. The LOD of the passive air nicotine badges was 0.005 ug/m³. We defined air nicotine concentrations greater than the LOD as positive for indoor SHS exposure. All methods are described in further detail in our previous work [12].

2.3. Statistical Analysis

Descriptive statistics were used to characterize the patient sample and the pollutant concentrations. Continuous variables were compared using Kruskal-Wallis tests. All analyses were performed using StataSE statistical software, version 12.0 (StataCorp, College Station, TX). Statistical significance was defined as p < 0.05.

3. Results

Ten homes were recruited and completed both visits. Half of homes used wood and/or coal for heating (**Table 1**). Overall PM concentrations were high (**Table 2(a)**). Over both seasons, median $PM_{2.5}$ concentrations were 34 ug/m³ in all homes, 61 ug/m³ in homes using wood and/or coal, 10 ug/m³ in homes with electric heat, and 52 ug/m³ in homes with gas heat. In the homes using wood and/or coal for heating, median $PM_{2.5}$ concentrations were 72 ug/m³ and 49 ug/m³ in August and December, respectively. Median $PM_{2.5-10}$ concentrations across seasons were 15 ug/m³ for all homes, 18 ug/m³ in wood/coal homes, 9 ug/m³ in electric homes, and 18 ug/m³ in homes with gas heat. In homes using wood and/or coal, August and December median $PM_{2.5-10}$ concentrations were 16 ug/m³ and 19 ug/m³, respectively.

Table 1. Housing characteristics.

	August n = 10	December n = 10
Primary heating fuel n (%)		
Electric	3 (30)	3 (30)
Coal only	1 (10)	1 (10)
Wood only	1 (10)	1 (10)
Wood and coal	3 (30)	3 (30)
Gas	2 (20)	2 (20)
Cooking stove fuel n (%)		
Electric	9 (90)	9 (90)
No stove	1 (10)	1 (10)
Smoking status n (%)		
Smoking in home	4 (40)	5 (50)
No smoking in home	6 (60)	5 (50)
General condition of home n (%)		
1-extremely poor	3 (30)	3 (30)
2-poor	1 (10)	1 (10)
3-average level of housekeeping	2 (20)	2 (20)
4-above average, clean	4 (40)	4 (40)
5-highly organized, extremely clean	0	0
Pets in home n (%)		
None	3 (30)	3 (30)
Dog	7 (70)	7 (70)
Cat	2 (20)	2 (20)
Bird	1 (10)	1 (10)

			(a)						
Results by Individual Home									
Primary heating fuel	Stove type	Smoker	PM _{2.5} (ug/m ³) August	PM _{2.5} (ug/m ³) December	PM _{2.5-10} (ug/m ³) August	PM _{2.5-10} (ug/m ³) December			
Wood + Coal	Electric	Yes	72.4	119.0	8.2	14.9			
Wood + Coal	Electric	Yes	133.8	152.6	12.5	88.0			
Coal	Electric	No	22.9	15.0	35.1	18.8			
Wood	Electric	Yes	49.9	34.3	16.0	16.2			
Wood	Electric	Yes	111.7	49.2	47.2	50.5			
Kerosene	No stove	Yes	68.9	81.8	16.4	19.9			
Propane	Electric	No	26.4	34.3	2.2	26.1			
Electric	Electric	No	10.0	7.0	9.0	14.4			
Electric	Electric	No	11.7	7.7	5.2	9.3			
Electric	Electric	No	10.4	12.8	5.8	10.8			
Results by Fuel Type Median (range)									
All homes $(n = 10)$			38.1 (10.0, 133.8)	34.3 (7.0, 152.6)	10.7 (2.2, 47.2)	17.5 (9.3, 88.0)			
Electric heat $(n = 3)$			10.4 (10.0, 11.7)	7.7 (7.0, 12.8)	5.8 (5.2, 9.0)	10.8 (9.3, 14.4)*			
Wood and/or coal heat $(n = 5)$			72.4 (22.9, 133.8)	49.2 (15.0, 152.6)	16.0 (8.2, 47.2)	18.8 (14.9, 88.0)			
Gas heat $(n = 2)$			47.7 (26.4, 69.0)	58.0 (34.3, 81.8)	9.3 (2.2, 16.4)	23.0 (19.9, 26.1)			

Table 2. (a) Pollutant concentrations: Particulate matter; (b) Pollutant concentrations: NO₂ and Nicotine.

BD = below limit of detection, *Denotes significant difference in pollutant concentration between August and December (p-value < 0.05).

			(b)						
Results by Individual Home									
Primary heating fuel	Stove type	Smoker	NO ₂ (ppb) August	NO ₂ (ppb) December	Nicotine (ug/m ³) August	Nicotine (ug/m ³) December			
Wood + Coal	Electric	Yes	2.2	2.2	2.39	7.96			
Wood + Coal	Electric	Yes	11.9	19.0	41.12	7.34			
Coal	Electric	No	3.6	20.2	0.96	0.45			
Wood	Electric	Yes	4.1	4.0	7.48	2.47			
Wood	Electric	Yes	3.4	2.6	11.84	1.52			
Kerosene	No stove	Yes	2.3	197.4	6.66	0.73			
Propane	Electric	No	1.6	124.0	BD	BD			
Electric	Electric	No	0.8	2.0	0.28	0.10			
Electric	Electric	No	1.4	2.1	BD	0.04			
Electric	Electric	No	1.7	1.8	0.94	0.07			
		Results b	y Fuel Type Media	an (range)					
All homes $(n = 10)$			2.3 (0.8, 11.9)	3.3 (1.8, 197.4)	1.7 (<0.01, 41.1)	0.6 (0.01, 8.0)			
Electric heat $(n = 3)$			1.4 (0.8, 1.7)	2.0 (1.8, 2.1)*	0.3 (<0.01, 0.9)	0.07 (0.04, 0.1)			
Wood and/or coal heat $(n = 5)$			3.6 (2.2, 11.9)	4.0 (2.2, 20.2)	7.5 (1.0, 41.1)	2.5 (0.4, 8.0)			
Gas heat (n=2)			2.0 (1.6, 2.3)	160.7 (124.0, 197.4)	3.33 (<0.01, 6.7)	0.4 (0.01, 0.7)			

BD = below limit of detection; *Denotes significant difference in pollutant concentration between August and December (p-value < 0.05).

Median NO₂ concentrations were higher in December (3.3 ppb) than August (2.3 ppb) with concentrations most notable in homes using gas fuel during the heating season (median concentration 161 ppb) (**Table 2(b**)). Fifty percent of participants reported smoking in their homes, and eighty percent of homes in August and 90% in December had detectable nicotine concentrations (**Table 2(b**)).

4. Discussion

We found high concentrations of indoor pollutants in our sample of homes in Central Appalachia, even during non-heating seasons. Concentrations of PM in our study homes were remarkably higher than levels reported in many US urban homes [13,14]. Although lower than many international settings, where biomass is used in open stoves for both cooking and heating [15-17], our sample had PM concentrations similar or high compared to other North American homes reporting regular wood and/or coal use [18-20]. For example, indoor PM2.5 concentration was less than 15 ug/m³ in a British Columbia population impacted by wood stoves [20]. A Rocky Mountain study evaluating the effect of woodstove replacement on indoor and ambient air quality reported a mean PM_{2.5} concentration of 51.2 ug/m³ prior to stove change out [18]. However, these studies and many other studies conducted in North American homes using wood and coal excluded smoking homes. In our population, given the high pollutant concentrations in the non-heating season and high prevalence of smoking, it is difficult to discern the relative contribution of wood/coal use and other particle-generating sources, like tobacco use, to the high PM concentrations found in this sample.

We did not find a statistically significant difference in PM concentrations between heating and non-heating seasons in homes using coal or wood for heat, but the coarse fraction of PM during the heating season was higher as compared to the non-heating season in homes using electric heat. These results were somewhat unexpected, but may represent the contribution of other non-heating particle-generating activities to PM concentrations and venting of coal/wood burning stoves to the outdoors. In addition, unmeasured differences between housing characteristics and the small sample size could partially explain the trends we observed.

Though difficult to make definitive conclusions based on the limited number of homes included in our sample, NO_2 concentrations in December were strikingly high in homes using gas space heaters, and remarkably higher than in urban environments where gas stoves are the predominant NO_2 source [21,22]. For example, the concentration of indoor NO_2 in Baltimore homes using gas stoves for cooking was 33.1 ppb, and similar results were found in Connecticut homes with gas stoves as the primary source of NO₂ [21,22]. In the Appalachia sample, mechanical characteristics of the space heaters, lack of sufficient ventilation, and smaller home size may contribute to the high concentration of NO₂ in these homes. Additional investigation is needed to characterize the impact of gas heaters and other combustion sources used for heating on indoor NO₂ concentrations.

To our knowledge, this is one of only two studies to report direct measurements of indoor air quality in rural homes in Central Appalachia [19]. Our results show the feasibility of indoor air monitoring in Central Appalachia. Furthermore, though our results do not definitively show an increase in pollutant burden during the heating season, they highlight the overall poor indoor air quality and high PM and NO₂ concentrations found in homes in this rural community. Given the known adverse health effects of pollutant exposure, further study is needed to investigate the contribution of poor indoor air quality associated with indoor combustion materials to negative public health outcomes seen in Appalachia, and to develop feasible interventions to help reduce the pollutant burden in the indoor environment.

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