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Assessment of Environmental Sources, Levels and Distribution of Polycyclic Aromatic Hydrocarbons within Nzoia Catchment Area in Kenya

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Abstract

Continuous concerns about Polycyclic Aromatic Hydrocarbons (PAHs) presence in the environment have raised concern because of their toxic effects to various organisms. Sugarcane farming and cane processing industries are major economic activities within River Nzoia catchment area in Kenya. For instance, the sugar industries produce wastes and by products which can cause PAHs emission and environmental contamination in addition to activities related to rapid urbanization that is being observed within the catchment. This study presents a report on sources and distribution of PAHs levels in sugarcane by products waste, sediments, water and soils within the River Nzoia catchment area. Soil and sediment samples were extracted by soxhlet extraction using dichloromethane and with C-18 catridges. Analyte separation and identification was done by GC-MS. Fourteen PAHs were detected with concentration ranges of; 0.6 μg/L - 80 μg/L for water, 0.01 μg/kg - 1200 μg/kg for soils and 0.13 μg/kg - 19.6 μg/kg for sediments. Bagasse waste had PAHs concentrations in the range of 0.4 - 14 µg/kg, and filter cake in the range of 1.7 - 30 μg/kg. Boiler waters reported the presence of 8 PAHs. The ratio of concentrations of PAHs in boiler water, filter cake and bagasse waste to the soils and water samples within the vicinity to the sugar processing companies did not indicate a point source of contamination; rather it pointed to diffuse sources. The same results were observed for water and sediment samples obtained in the vicinity of waste dumpsite. Variation of PAHs concentrations from sugar manufacturing processes corresponded to the kind and conditions of the processes. Lower molecular weight PAHs dominated in

water sample. The presense of benz: 1) pyrene, benz 2) flourancene and Indeno(123,cd)pyrene in both water and sludge soils are of concern since this water is abstracted for domestic use, while sludge soil is used as fertilizer in agricultural farms. Although the levels of PAHs obtained in this study were below the established environment and human health safefty limits, the results underscore the need for mornitoring levels and determining potential sources for PAHs in the environment.

Keywords

Polycyclic Aromatic Hydrocarbons, Sugarcane by Products, Sources and Distribution, Waste Disposal

1. Introduction

The term polycyclic aromatic hydrocarbons (PAHs) refers to organic chemical compounds that consist of two or more fused aromatic rings and do not contain heteroatom or carry substituents [1]. PAHs are regarded as environmental pollutants [2] [3] that are formed mainly due to incomplete combustion of waste [4] [5]. Other anthropogenic sources of PAHs include vehicle exhaust, coal combustion for cooking and heating, industrial effluent, domestic effluent and municipal waste incineration among others [6]. The United States environmental protection agency (US EPA) prioritised sixteen PAHs of which, those with high molecular weight PAHs are of concern because they have been identified as potential carcinogens [7], mutagens and teratogens [8]. Therefore, there is need to ascertain the sources and distribution of PAHs in places where there is potential of occurrence and contamination in industrial processes such as in sugarcane production and processing [9]. Studies in Brazil reported the presence of benzo[a]anthracene (BaA), benzo[a]pyrene (BaP), and dibenzo[a,h]anthracene (DBA) among other PAHs in cachacüa, (a Brazilian sugar cane spirit), where it was observed that samples produced from burned sugar cane fields had higher PAH levels than those produced from non-burned sugar cane [10]. Therefore sugarcane burning was found to be a source of PAHs. Sugar cane by products such as filter mud, bagasse, boiler water and molasses [11], may contain PAHs. Nzoia catchment area is the main sugar belt region of Kenya. The catchment is a home for sugar industries, for instance Mumias, Butali, West Kenya and Nzoia sugar companies. There is need to assess the levels, sources and distribution of PAHs among the by products of these sugarcane industries, and also their distribution in sediments, water and soils within the region. It is expected that the agricultural soils around the sugar factories and urban areas in the study region will have higher concentrations of PAHs than the ones further away from these areas.

Sediments act as a sink for PAHs [12], which may originate from urban centres [13]. Lack of sufficient data on PAHs pollution in soil, river water and sedi-

ments environmental pollutants is an obstacle to put in place mitigative measures for controlling PAHs pollution, hence the need for this study. Levels sources, distribution, occurrence, transport, and fate of PAHs have been documented in various matrices samples [14] [15] [16] and compared in both the urban areas rather than those of rural areas [17] [18]. In Kenya, levels of PAHs in water and sediments [19] [20] [21] [22] [23], burning fuel biomass [24], urban air [5] have been previously reported. PAHs analysed in soil, water, sediments and other matrices can provide important information on the state of environmental pollution and the concentrations of individual PAHs in soil resulting mainly from motor vehicle exhaust that usually ranges between 1 and 2000 µg/kg [25]. This study was carried out in order to assess the PAHs environmental levels and their occurance in sugar production process waste, as well as determine their Distribution within Nzoia Catchment Area.

2. Materials and Methods

2.1. Study Area

This study was conducted in the sugar belt regions of the Nzoia River catchment area of Kenya. River Nzoia Catchment lies between latitudes 1°30'N and 0°05'S and longitudes 34° and 35°45'E as shown in Figure 1. One of its major tributaries of the River Nzoia is River Isiukhu. The sampling locations within Nzoia River Catchment were selected depending on the social-economic activities and waste generation and disposal methods within the study location. The specific sampling locations were; Shirere, Rostaman, Mwibatsiro, Rostamandumpsite, Mmust dumpsite, Shirere, Shibeye, Ekero, West Kenya sugar company, Nzoia sugar company, Mumias sugar company and within Kakamega Municipality. Soil samples were collected from Mumias farm, Mumias nuclear estate. Sludge soil was collected from the Municipal dumpsite and Mmust dumpsite. Sediment and water samples were collected from Mumias sugar company stream, at Ekero, Shibeye and Rosterman all along Isiukhu river, Mwibatsiro and MMUST stream. Within the sugar company, the by products as waste from the sugar production proceses sampled for analysis were; Baggasse waste from Nzoia Sugar Company (NSC), Baggasse waste from West Kenya Sugar Company (WKSC), Filter cake WKSC, Filter cake NSC, Boiler water NSC and Boiler water WKSC. Table 1 describes the area where samples were collected.

The selection of these points was based upon human activities taking place near the river banks such as sand harvesting, farming, waste and effluents disposal methods and accessibility. The area surrounding the sampling sites had sugarcane farming as the main economic activity. The sampling sites, rivers, and the forest within the study area are shown in **Figure 1**.

2.2. Chemicals and Apparatus

High purity standard for PAHs was purchased from Sigma Aldrich, UK. Solid Phase Extraction cartridges and nylon micro filters were purchased from Merck

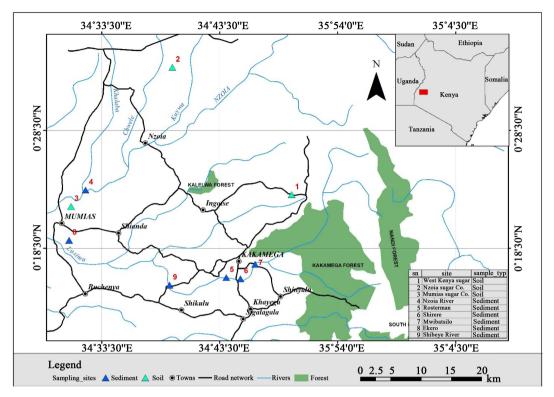


Figure 1. Map of River Nzoia catchment area showing the study locations (numbered 1 to 13). WKSC (1), NSC (2), Mumias sugar company stream (3), Mumias Nuclear Estate (4), Rosterman (5), Shirere (6), Mwibatsiro (7), Ekero (8), Shibeye (9), MMUST (10), Municipal dumpsite (11), MMUST stream (12), Municipal sludge soil (13).

Table 1. Description and coordinates of sampling locations.

Sampling site	Type of sample	Coordinates of the site	Location description
Shirere	Sediment and water	34°44'30"E 0°28'30"N	Sand harvesting and farming along the river.
Rostaman	Sediment and water	34°43'30"E 0°18'N	Gold mining, informal settlements, illegal wastes and discharges damping site due to mining activity and maize farming along the river banks.
Mwibatsiro stream	Sediment and water	34°54'0"E 0°18'N	Maize and Napier grass farming along the river
Municipal dumpsite	Soil	34°53'20"E 0°28'30"N	Kakamega town damping site; industrial, agricultural and domestic wastes from the town and the neighborhood localities.
Shirere sewage	Soil, sediments and water	34°45'0"E 0°28'30"N	Sand harvesting and maize farming along the river.
Ekero	Sediments and water	34°33'0"E 0°28'30"N	Sugarcane farming, and maize farming.
MMUST damp site	Soil	34°29'20"E 0°18'N	Main MMUST dumpsite, kitchen wastes and domestic wastes.
Mayoni	Sediment and water	34°43'28"E 0°18'N	Sugarcane farming, sand harvesting and car wash along the river banks.
Shibeye	Sediments and water	34°33'20"E 0°28'30"N	Sugar cane farming.
West Kenya sugar Company	Boiler water and mud filter	35°4'30"E 0°18'N	Sugar processing, accumulation and transportation of mud filter to various farms.
Mumias sugar Company	Burnt farm soil and nuclear estate soil	34°32'20"E 0°18'N	Sugar processing, accumulation and transportation of mud filter to various farms.
Nzoia sugar Company	Bagasse soil, mud filter and boiler water	34°46'30"E 0°28'N	Sugar processing, accumulation and transportation of mud filter to various farms.

Chemicals. Anhydrous Sodium sulfate (Na_2SO_4), Dichloromethane and Iso-octane were of analytical grade reagent purchased from Kobian Chemicals Nairobi. Internal standard dodecane was purchased from Sigma Aldrich, UK and Reference standards of PAHS obtained from IoIc (Warsaw: Puder, Poland). Soxhlet extraction apparatus, SPE cartridges, 0.45 μ m PTFE (polytetrafluoroethylene) syringe filter, cellulose extraction thimble, from Kobian Chemicals Nairobi and GC-MS instruments with the following specifications: GC Model of 7890A manufactured by Agilent technologies, Inc, Beijing China, with a MASS detector of 5973C by Agilent Technologies, Inc. Santa Clara, CA, USA attached to an auto sampler of 7683 by Agilent Technologies.

2.3. Sampling

Reconnaissance study was done to obtain information regarding the locations where sampling was done, based on the activities that take place within the locations (see Table 1). Sample was collected from sites believed to have previous history of anthropogenic activities that include open burning or combustion such as dump sites, sugarcane fields, urban areas, rivers in the vicinity of sugar factories, and also in sites that have no history of interferences by anthropogenic activities such as Kakamega forest, for comparison. Control samples were obtained within the Kakamega forest. Sediment sample was collected using an Ekman Grabber. Water sample was collected in amber bottles and soil samples were collected using a stainless steel hoe and scooped using stainless steel shovel. The samples were then wrapped in clean piece of aluminum foil, packed in black polythene bag, and labeled. The entire sample was placed in cooler boxes and transported to the university laboratory where they were stored in the refrigerator at 4°C.

2.4. Water Sample Extraction and Analysis

Extraction of water samples was done by solid phase extraction procedure which was adopted from US EPA Method 3510C [26]. 500 mls of water sample was measured using a glass measuring cylinder, transferred into a 1.0 L beaker and its pH recorded. The water sample was filtered using what man filter paper No.1 to remove other impurities. The pH was adjusted by adding drops of 0.1 M hydrochloric acid or 0.1 M sodium hydroxide solutions with carefully stirring to adjust the pH to 7.0. The neutral solution was transferred to 500 ml beaker and treated with 25 g of sodium chloride to salt out the PAHs from aqueous phase to organic phase. The mixture was further shaken gently and allowed to settle for 30 minutes to enhance homogeneity. Extraction process was repeated thrice each time with 30 ml dichloromethane. The combined extracts were dried using activated anhydrous Na₂SO₄. 2 ml of isooctane was added as keeper of the PAHs then concentrated to about 2 ml. The concentrated extracts were filtered with 0.45 μm PTFE (polytetrafluoroethylene) syringe filter, and the volume adjusted to the 2 mls mark using the mobile phase solvent before the extract was stored in

a fridge at -4 °C prior to GC-MS analysis [27]. Dodecane was used as the internal standard.

2.5. Sediment and Soil Samples Extraction and Analysis

Sediment samples were removed from freezer and allowed to thaw for about 4 hours prior to extraction. Samples were air-dried in the dark, crushed and sieved with a 2 mm sieve. About 10 grams of the sieved samples were weight and put into the Soxhlet extraction thimble and 0.1 of 0.01 μ l/ml dodecane solution added as the internal standard. This was extracted with 250 ml of dichloromethane in a round bottomed flask for at least 8 hours in a Soxhlet extractor set-up. The extracts were dried using 2 g of anhydrous sodium sulphate. 2 ml of isooctane was added as keeper, and the mixture concentrated to 2 ml. The concentrated extracts were filtered through a 0.45 μ m PTFE (polytetrafluoroethylene) syringe filter, stored in a fridge at -4 °C prior to GC-MS analysis.

2.6. Instrumental Sample Analysis

The samples were run using GC-MS: GC Model of 7890A hyphenated with a MASS detector of 5973C by Agilent Technologies, Inc. Santa Clara, CA, and USA attached to an auto sampler of 7683 by Agilent Technologies. The GC was equipped with a DB-5 silica fused high performance capillary column, (30 m length, 0.25 mm internal diameter and 0.25 µm film thickness). The carrier gas was helium (99.99%) with a constant flow rate of 2 ml/min. Make up gas was (99.99%) with a constant flow rate of 2 ml/min. The injection volume was 1 ul with a pulsed split less injection mode. The oven temperature was started initially at 80°C where it was held for 3 mins. Programmed for two temperature ramps was from 80°C to 280°C with increase at 10°C·min⁻¹ and held for 20 mins. Injector temperatures were 250 and detector temparature was 300°C. Solvent delay was programmed at 5 mins. The mass selective detector was operated in the electron ionization (70 eV) and the selected ion monitoring modes. Mass range was between m/z 30 - 450. Samples for GC-MS analysis were evaporated to dryness under reduced pressure and residue re-dissolved in 50 µL of iso-octane. Analysis by GC-MS was performed in split/splitles mode. Mass spectrometry was operated on electron impact ionization mode. Analyte identification was also based on the comparison of mass spectra of unknowns with those in the NIST/EPA/NIH MASS SPECTRAL LIBRARY library [28]. Quantitative analysis of PAHs was achieved through the integration of selected GC chromatograms.

2.7. Analytical Quality Control

Extraction efficiency was tested by obtaining recovery values. Five blank samples which were prepared in the same way that the real analytical samples were prepared were used for each set of ten real analytical samples analysed. Calibration curves were drawn for all the elements to be analyzed and regression equations

determined for every calibration curve, to check the instrumental response to various analyte concentrations. In all the cases regression was found have $r^2 = 0.99$ and above. Accuracy and precision of the method were also tested.

3. Results and Discussion

Analytical quality control included determination of compound-specific estimating instrument detection limits (IDL) and method detection limits (MDL) and the calculation of internal standard recoveries. Recovery values were obtained by spiking each matrix with internal standard and ranged from 61,23% to 72.89%. There were no detected PAHs in the blank samples. The IDL was set at a signal to noise ratio of 3, as obtained from the chromatograms and the MDL was defined as the mean concentration of the compounds plus 3 times the standard deviation of signals obtained from the procedural blank experiments. The IDL and the MDL are shown in Table 2. Control samples were obtained from stream and soil within, deep in the Kakamega tropical rain forest. The concentrations obtained from the water control samples were 0.06 µg/L for naphthalene and 1.10 µg/L for pyrene. The other PAHs gave concentarion values below the MDL. Concentrations obtained from the control soil sample were 1.1 µg/kg for anthracene and 0.08µg/kg. Other PAHs in control soil samples had their concentrations ≥ MDL. The linearity response obtained from the calibration standard as shown in Table 2 and was found to be good. Accuracy and precision of analyses were found to be satisfactory. Table 2 also shows the molecular ion that was used to identify the PAHs in the chromatograms.

Table 2. PAHs analyzed showing the m/z monitored, the standard linearity response r² values, the percentage (%) recoveries and the IDL and MDL set values.

Peak no.	PAHs	m/z	Correl. Coeff.	Recovery (%)	IDL	MDL
1	Naphthalene (Nap)	128/130	0.9953	72.89	0.02	0.06
2	Acenaphthylene (Acy)	152/154	0.9953	70.64	0.02	0.5
3	Acenaphthene (Ace)	153/154	0.9968	64.73	0.02	1.0
4	Fluorene (Flu)	166/168	0.9965	71.12	0.2	0.5
5	Phenanthrene (Phe)	176/178	0.9985	61.62	0.2	1.0
6	Anthracene (Ant)	176/178	0.9956	71.25	0.05	0.5
7	Flourancene (Flr)	200/202	0.9974	72.64	0.2	1.0
8	Pyrene (Pyr)	200/202	0.9968	64.94	0.05	1.0
9	Benz(a)anthracene (BaA)	226/228	0.9964	71.34	0.05	1.0
10	Chrysene (Chr)	226/228	0.9962	62.61	0.05	0.2
11	Benz(b)flourancene (BbF)	250/252	0.9960	70.64	0.05	0.1
12	Benz(a)pyrene (BaP)	250/252	0.9943	61.84	0.05	0.2
13	Indeno(123,cd)pyrene (InP)	276/278	0.9960	61.23	0.05	0.1
14	Dibenzo(def,mno)chrysene (DBC)	276/278	0.9979	64.78	0.05	0.1

Detection of PAHs varied from one sampling point to another. For example, fluorene (Flu) and chrysene (Chr), were not detected in all the sampling locations within the sugarbelt region, while naphthalene (Nap) was detected in almost all samples analysed. The results of the concentration of PAHs obtained for water, soil and sediment samples are shown in Table 3. Kakamega municipal dumpsite samples had the highest concentration obtained for Nap, 160 µg/kg and for Anthracene (Ant) 1200 µg/kg. The results obtained at the Kakamega municipal dumpsite were different to those obtained at MMUST dumpsite where the concentration of Nap was 6.0 µg/kg. The difference in the amount of waste between the two dumpsites could be the probable reason to explan the difference. While Kakamega dumpsite services a population of about half a million people, the MMUST dumpsite is community owned and services about five thousand people. Concentrations of acenaphthene (Ace), flourancene (Flr), chrysene (Chr) and Dibenzo(def,mno)chrysene (DBC) were below MDL. It was interesting to note that samples of municipal sludge soils had concentrations of 3.80 µg/kg for Ant, 8.0 µg/kg for Pyrene(Pyr), 1.72 µg/kg for Benz(a)pyrene (BaP) and 1.12 μg/kg for Indeno(123,cd)pyrene (InP). The addition of sewage sludge to agricultural soils can be a source of PAHs in these soils. PAHs concentrationis proportional to the amount of sewage sludge applied to agricultural field soils [29]. This is interesting considering that sewage sludge is often used as fertilizer.

Table 3. Concentrations of the 16 priority PAHs in different matrices in different sampling sites (Soils: µg/kg and Water: µg/L).

Matrices	Sampling location	Nap	Acy	Ace	Flu	Phe	Ant	Flr	Pyr	BaA	Chr	BbF	BaP	InP	DBC
	Mumias farm	1.32	0.60	-	0.80	6.00	-	-	-	3.4	-	-	-	-	-
	Mumias nuclear estate	1.50	0.80	-	0.88	6.0	20	-	-	-	-	-	-	-	-
Soil	Kakamega municipal dumpsite	160.0	-	-	-	-	1200	-	-	-	-	10	-	-	-
	Municipal sludge soil	-	-	-	-	-	3.8	-	8.0	-	-	-	1.72	1.12	-
	MMUST dumpsite	6.0	-	-	-	-	-	-	-	-	-	-	-	4.0	-
	Mumias sugar Co. stream	6.0	-	-	5.0	-	-	-	-	-	1.20	2.0	-	-	4.0.
	Ekero	32.0	-	-	14	-	80.0	-	-	-	-	4.0	-	-	-
Sediment	Shibeye	4.50	-	-	0.94	-	40.0	-	-	-	-	-	-	-	-
Sediment	Rosterman	-	-	-	-	-	-	-	-	-	-	1.14	-	-	0.14
	Mwibatsiro	19.60	-	-	5.2	-	6.0	-	-	-	-	-	-	-	-
	MMUST stream	-	-	-	-	-	-	-	-	-	-	0.14	-	0.12	-
	Mumias sugar Co. stream	15.60	-	-	-	-	-	-	-	-	-	-	5.2	-	-
	Ekero	0.12	3.60	-	-	-	-	-	3.16	-	-	-	-	-	-
Water	Shibeye	0.08	-	-	-	3.6	0.08	-	-	-	-	-	6.8	-	-
water	Rosterman	0.06	-	11.6	-	-	-	-	1.20	-	-	4.80	-	-	-
	Mwibatsiro	6.60	-	-	-	-	-	-	1.10	-	-	4.72	-	-	-
	MMUST stream	38.4	-	-	-	-	-	-	0.72	-	-	-	-	-	-

⁻ means method detection limit (<MDL). Naphthalene (Nap), Acenaphthylene (Acy), Acenaphthene (Ace), Fluorene (Flu), Phenanthrene (Phe), Anthracene (Ant), Flourancene (Flr), Pyrene (pyr), Benz(a)anthracene (BaA), Chrysene (Chr), Benz(b)flourancene (BbF), Benz(a)pyrene (BaP), Indeno(123,cd)pyrene (InP), Dibenzo(def,mno)chrysene (DBC).

The ratios of concentrations of PAHs from sugarcane processes samples to those of water and sediment samples in the vicinity of the processing companies indicated a diffuse source of the PAHs rather than specific point source.

The concentration obtained for naphthalene (Nap), Fluorene (Flu), Anthracene (Ant), Chrysene (Chr), Benz(b)flourancene (BbF), Benz(a)pyrene (BaP), Indeno(123,cd)pyrene (InP), Dibenzo(def,mno)chrysene (DBC) ranged as follows; 4.50 - 19.60 µg/kg, 0.94 - 14.0 µg/kg, 6.0 - 80.0 µg/kg, 0.14 - 4.0 µg/kg, 0.12 µg/kg, 0.14 - 4.0 µg/kg in all sampling locations respectively. Sediments obtained from Ekero had the highest concentrations of Ant (80.0 µg/kg) and contained five PAHs analysed. Ekero location has intense sugarcane and maize farming. Nap was present in all water samples analysed with the highest concentration of 15.60 µg/L. The Mumias sugar company stream carries effluent from and around the sugar factory vicinity. The results indicated that surface runoff from the factory farms and waste from the sugar cane processing may be contributing to the presence of the PAHs observed in this sampling location.

Samples of waste and by products from three sugar processing factories namely Nzoia sugar company (NSC) and West Kenya sugar company (WKSC) were analysed for the PAHs. Bagasse waste samples from NSC had concentrations of Nap, Acy, Flu, Ant and InP ranging from 0.4 to 14 µg/kg, while filter cake from the same factory had concentrations of Nap, Phe, Ant, Flr, Inp and DBC ranging from 1.7 to 14 μ g/kg (see **Table 4**). Comparing the two companies, bagasse waste from WKSC had only Nap quantified with a concentration of 2.0 µg/kg. However, the filter cake sample from WKSC had quantifiable levels of Flu, Phe, Ant, Flr, InP and DBC in the concentration range of 5.2 to 30.0 µg/kg. The variation of PAHs concentrations in waste and by products from sugar manufacturing processes is related to the kind and conditions of the processes. The same conclusion can be made from the concentrations of PAHs obtained in boiler water in NSC and WKSC where Pyr, BbF, InP, DBC and Nap, Ant, Pyr, BaA were quantified respectively. Table 4 shows the concentrations of PAHs obtained from sugarcane process waste from NSC and WKSC. Previous studies have indicated that milling process can course contamination [11] along the process hence leading to PAHs introduction to the process.

Table 4. Concentrations of PAHs in different sugar process by products and wastes analysed. Concentration in solid waste is in $\mu g/kg$ and water in $\mu g/L$.

Matrices	Nap	Acy	Ace	Flu	Phe	Ant	Flr	Pyr	BaA	Chry	BbF	BaP	InP	DBC
Bagasse waste NSC	12.0	8.0	-	0.60	-	14.0	-	-	-	-	-	-	0.4	-
Bagasse waste WKSC	2.0	-	-	-	-	-	-	-	-	-	-	-	-	-
Filter cake WKSC	-	-	-	30.0	5.20	14	7.0	-	-	-	-	-	30.0	28.0
Filter cake NSC	1.80	-	-	-	3.20	8.0	14.0	-	-	-	-	-	1.7	12.0
Boiler water NSC	-	-	-	-	-	-	-	1.20	-	-	0.12	-	0.12	0.08
Boiler water WKSC	0.07	-	-	-	-	0.10	-	1.28	12.0	-	-	-	-	-

Naphthalene (Nap), Acenaphthylene (Acy), Acenaphthene (Ace), Fluorene (Flu), Phenanthrene (Phe), Anthracene (Ant), Flourancene (Flr), Pyrene (Pyr), Benz(a)anthracene (BaA), Chrysene (Chr), Benz(b)flourancene (BbF), Benz(a)pyrene (BaP), Indeno(123,cd)pyrene (InP), Dibenzo(def,mno)chrysene (DBC). Means below limit of detection (<MDL).

Various industrial manufacturing processes may induce the generation of PAHs. For example, [30] assessed the contamination of rapeseed oil by PAHs in processes that require temperature control. In the study, it was reported that Benzo[a]pyrene, Benzo[a]fluoranthene, Benzo[b]fluoranthene and chrysene determined in rapeseeds as a result of processing. [31] reported the presence of 16 PAHs in fresh tea leaves due to manufacturing process. In the study, Benzo(a)pyrene which is classified as a 2A class carcinogen was found in the dried tea leaves only whereas, naphthalene which is classified as a 2B probable human carcinogen was present during all the stages of manufacturing.

3.1. Distribution of PAHs in Different Sample Matrices in the River Nzoia Catchment Area

The distribution of PAHs varied among the sampling locations and the sample matrices. There was no single sample where all the priority PAHs were detected. Variation in concentrations indicated different sources of PAHs [32]. The observed concentrations distribution in the Nzoia catchment region, indicates varied sources of the studied PAHs. Figure 2 shows the distribution in percentage

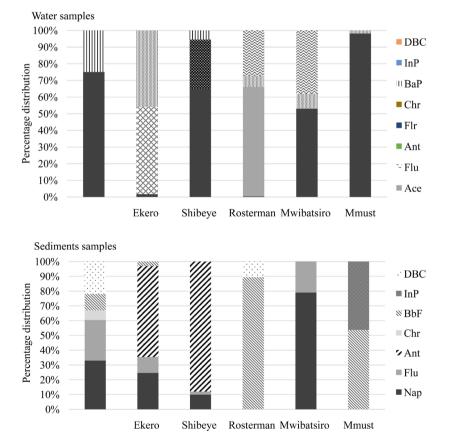


Figure 2. Percentage (%) of distributions of PAHs contribution in water and sediment samples within the study locations. Naphthalene (Nap), Acenaphthylene (Acy), Acenaphthene (Ace), Fluorene (Flu), Phenanthrene (Phe), Anthracene (Ant), Flourancene (Flr), Pyrene (Pyr), Benz(a)anthracene (BaA), Chrysene (Chr), Benz(b)flourancene (BbF), Benz(a)pyrene (BaP), Indeno(123,cd)pyrene (InP), Dibenzo(def,mno)chrysene (DBC).

(%) of PAHs in water and sediments samples. Mumias sugar company stream sample had only two PAHs present with Nap distribution in the sample at over 75%. Ekero, Shibeye, Mwibatsiro and Rosterman which are located along Isiukhu river had at least three PAHs in the water samples.

However, Acy, Nap and Ace make up to more than 50% of total PAHs in water of the three locations respectively. Benz(a)anthracene (BaA) is dominat in water sample from MMUST stream (over 90%) and Mwibatsiro (Over 50%) (see Figure 2). In a study by [33] that BaA was among the PAHs that showed a good correlation indicating that they originate from the burning processes. Classification of PAHs is based on molecular weights (MW). The classes of PAHs strongly correlate with their environmental fate. Lower molecular weight is PAHs of 152 to 178 g/mol. The medium (middle) molecular weight PAHs, have about 202 g/mol and High Molecular weights have MW of 228 to 278 g/mol. Predominance of low and medium molecular weight PAHs in water and sediment samples reflected a similar source. The high MW class of PAHs has low solubility in water. These could explain the observed low or/and below method detection limit concentrations of BaA, Char, BbF, BaP, InP and DBC in water samples. The possible contributors of PAHs determined in sediment samples might be through surface runoffs from wastes dumping sites along the river banks. Discharge sewage effluents from the sewerage plant may be contributing significantly to Nap and Flu present in the Mwibatsiro sediments. High hydrophobicity of PAHs is associated with their adsorption on sediments, organic matter and suspended particles [34]. Naphthalene and Phenanthrene are degraded rapidly in sediments but higher MW PAHs such as Benzo(a)pyrene are more recalcitrant [35].

Soils are likely to retain high MW PAHs. The EPA has classified Benz(a)anthracene (BaA), Benz(a)pyrene (BaP), Benz(b)flourancene (BbF), chrysene (Chr), Dibenz(ah)anthracene, and Indeno(1,2,3-cd)pyrene (InP), as probable human carcinogens [7]. Of these carcinogenic PAHs, Benz(a)pyrene (BaP), Benz(b)flourancene (BbF) and Indeno(123,cd)pyrene (InP), were major contributors to the river water and sludge soil. This observation is of major concern since water from the rivers within the study region is abstracted for domestic use. Also, the sludge soil is used as fertilizer in agricultural farms (Figure 3). Show the percentage of distributions of individual PAHs in soil samples obtained from different localities within the Nzoia River catchment area. Naphthalene (Nap), was distributed in all soil samples except the sludge soil sample. It was higher in the MMUST sample (60% distributed). The MMUST soil sample also contained Indeno(123,cd)pyrene (InP), which tends to adsorption strong onto the soil organic matter. Anthracene recorded the highest concentration of 20 μg/kg and fluorine with the least concentration level of 0.8 μg/kg in Mumias nuclear farm. Observed concentrations of PAHs at this sampling location can be attributed to agricultural activities such as cane farming within the study area is associated with cane burning, during land preparation for another crop. At the

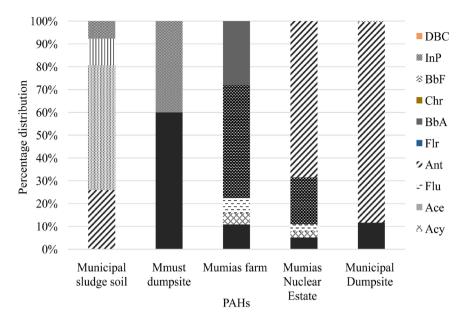


Figure 3. Percentage (%) of distributions of PAHs contribution in soil samples within the study locations.

Kakamega municipal dumpsite, the high concentration of anthracene of 1200 μg/kg and Benzo(b)flourancene may therefore be as a result of open burning, which was the observed method used to reduce the waste volumes at dumpsites within the study region. The distribution of anthracene in soil samples was over 85%, 65% and 25% in municipal dumpsite, mumias nuclear estate and municipal sludge soil samples respectively (see **Figure 3**). Anthracene is to be one of the most prevalent PAHs in soils in the study region. PAHs are generated by incomplete combustion, thermal means such as fires and burning or by other anthropogenic sources [36] [37]. Higher concentrations of PAHs in urban soils than in rural soils is a clear indicator of urban centers associated anthropogenic activities contribution of PAHs in the environment.

Anthracene was quantified at the municipal dumpsite but not in water and sediment within the dumpsite area. This observation indicated a diffuse source of anthracene. Naphthalene with a concentration of 0.06 $\mu g/L$ and 160 $\mu g/kg$ within the vicinity of munipal dumpsite at Rosterman could indicate that the source of naphthalene in water was the municipal dumpsite because of the ratio of 160:0.06 for the dumpsite and water respectively. However concentration of naphthalene at MMUST dumpsite was 6.0 $\mu g/kg$ and was lower than the concentration obtained in water within the same vicinity which was 38.4 $\mu g/L$ (ratio of 0.06:38.4) an indication of other various sources of naphthalene in water.

Decomposition of waste at dumpsites, other pyrogenic processes and open burning that take place at the municipal dumpsite may produce and/or release PAHs to the soil. Phenanthrene was distributed in about 50% of the mumias nuclear estate farm and about 20% of the soils from municipal dumpsite respectively, even though phenanthrene is espected to be removed from soils due to its relatively high water solubility.

3.2. Some International Data on Concentrations of PAHs in Various Matrices

Polycyclic aromatic hydrocarbons are ubiquitous in the modern environment [38]. Results from around the globe from various researchers generally show a variation in concentrations of PAHs among countries, and sampling locations. [21] researched on PAHs in water and sediments of car wash and kisat areas of winam Golf of Lake Victoria in Kenya. The PAH concentrations in sediment and water ranged from 0.04 to 31.95 µg/kg dry weight and 3.32 to 55.8 µg/L, respectively, depending upon the sampling location. [39] studied the distributions of PAHs in surface waters, sediments and soils of Hangzhou city, China and reported that the total concentrations of 10 PAHs in water ranging from 0.989 μg/Lto 9.663 μg/L, with a mean concentration of 3.717 μg/L. The composition pattern of PAHs in the water showed that PAHs (naphthalene, fluorene, phenanthrene, and pyrene) were the most abundant in all water samples. Five-ring PAHs Benzo(a)anthracene, Benzo(b)fluoranthene, chrysene, and Benzo(a)pyrene concentrations gave the lowest concentrations. Additionally, another study reported in Niger Delta, Nigeria where PAHs concentrations were averaged at 0.1761 mg/L was determined in selected water bodies in Niger Delta and 1.95 μg/L to 10.9 μg/L determined in Ekpan Creek [40]. In Ghana, total PAHs reported their concentration range up to 14.587 µg/L [41]. Another study by [42] from China reported that areas with intense navigation and large industrial facilities including oil refineries had higher concentration of PAHs in sediments. In the study, the range of PAHs in concentration obtained ranged from 464 to 2.62 μg/kg as shown in **Table 5**.

Study done in China reported PAHs levels from 0.016 μ g/kg in rural to 3.884 μ g/kg in the suburban soils of Beijing, China [43]. These levels are higher than those obtained from this study. According to [44] observed that industrial sites had higher PAHs levels than other more remote regions studied. Range in concentrations of PAHs determined by Oketola and [45], in the leachate samples in Lagos Nigeria was 0.85 to 1.47 mg/L and for top soil around the dumpsite, the values 0.94 - 2.79 mg/kg. Poor management of dumpsites can cause the leaching of leachate to ground water and surface water through run-off, and can have adverse effects on human health and the ecosystem. [46] obtained high concentrations of total PAHs, inside the landfill (1475 μ g·kg⁻¹ dw), whereas concentrations in the surrounding soils were by far lower ranging between 11.2 - 28.1 μ g·kg⁻¹ dw for PAHs.

4. Conclusions and Recommendations

The environmental levels, sources and distribution of the sixteen priority PAHs were reported for the first time within the different matrices in the Nzoia River catchment area. Sewerage system could also be a potential pathway of PAHs to the environment. Results of PAHs concentrations in sludge soils that originate from the wastewater treatment plant within Kakamega town (the municipal

Table 5. International data on concentrations of PAHs in various matrices studied, compared to data from this study.

Type of sample	Concentrations range	Country	Place of study	Reference
	≤MDL - 1200 μg/kg	Kenya	Nzoia catchment area	This study
	0.85 - 1.47 mg/L	Nigeria	Lagos	[45]
Dumpsite and Landfill	$1475~\mu g\cdot kg^{-1}~dw$	Greece	Nothern Greece	[46]
	9.9 - 575.6 ng/g dw	Malaysia	Taman Beringin	[47]
	≤MDL - 8 µg/kg	Kenya	Nzoia catchment area	This study
	0.01 - 3.884 μg/kg	China	Sub urban soils	[43]
	3.1 - 28.5 μg/kg	India	Industrial sites	[44]
Soil	2.0 - 6 μg/kg	France	Highway soils	[48]
	0.94 - 2.79 mg/kg	Nigeria	Lagos	[45]
	11.2 - 28.1 μg/kg	Greece	Nothern Greece	[46]
	≤MDL - 80 μg/kg	Kenya	Nzoia catchment area	This study
	37.7 - 11.78 μg/kg	Brazil	Patos Lagoon	[49]
	464 - 2.621 μg/kg	China	Industrial facilities	[42]
Sediment	0.3 - 13.5 μg/kg	Kenya	Car wash, Kisat, Kisumu	[50]
	0.04 to 31.95 μg/kg	Kenya	Winam Gulf	[21]
	$1.5 \times 10^{-5} \text{ ng/g}$	Malaysia	Taman Beringin	[47]
	1107 - 22,310 μg/kg	South Africa	Buffalo River Estuary	[51]
	≤MDL - 38.4 μg/L	Kenya	Nzoia catchment area	This study
	0.989 - 9.663 μg/L	China	Hangzhou city surface waters	[39]
TAT .	0.0 - 14.58 μg/L	Ghana	Surface water	[41]
Water	10 - 20 ng/l	Poland	IDLz river	[52]
	1.95 - 10.5 μg/L	Nigeria	Niger Delta	[40]
	14.91 - 206 μg/L	South Africa	Buffalo River Estuary	[51]
	3.32 -55.8 μg/L	Kenya	Winam Gulf	[21]

sludge soil sample) had Ant, Pyr, BaP and InP in the range of 1.12 - $8.0~\mu g/kg$. This observation indicated that poor waste management contributes significantly to the introduction of PAHs to the respective segments of the environment. Filter cake, bagasse and boiler water samples obtained from two sugar industries contained PAHs. Surface runoff from the factory farms and waste from the sugar cane processing processes may be contributing to the presence of the PAHs. This study recommends the use of cleaner operational conditions by sugar companies, and routine analysis of waste and by products from the processes geared towards eradication of emission and contamination through the processes. Results from these study point towards a variety of sources of PAHs, which can be biomass burning during land preparation, waste reduction by burning at waste dumpsites, and the numerous anthropogenic activities within the urban settle-

ments within the Nzoia River catchment area. The predominance of low and medium molecular weight PAHs in water and sediment samples was observed, which can be attributed to the low solubility of the higher MW class PAHs. High hydrophobicity of PAHs is also associated with their presence on sediments and suspended particles. Potential carcinogenic PAHs such as Benz(a)pyrene (BaP), Benz(b)flourancene (BbF) and Indeno(123,cd)pyrene (InP) were quantified in river water and sludge soil, which is a major pathway of these compounds to aquatic organisms and human. Mitigation measures that are informed by routine analysis of soil and water samples to identify if there are hotspots soursces of PAHs within the catchment are recommended.

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Conflicts of Interest

The authors declare no conflicts of interest regarding the publication of this paper.

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