

Emerging Water Quality Issues along Rio de la Sabana, Mexico

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Abstract

The basin of Rio de la Sabana is the largest tributary of the Tres Palos coastal lagoon in Southwest Mexico, east of Acapulco. This lagoon and its upstream basin areas have become a high priority area for the preservation of coastal and marine environments. To obtain information about water quality as affected by urban expansion since 2002, fourteen physicochemical parameters (temperature, pH, electrical conductivity, dissolved oxygen, ammonium, nitrate, nitrite, sulphate, phosphate), biochemical (biological and chemical oxygen demand, methylene blue active substances) and bacteriological parameters (total and fecal coliforms) were determined. This sampling was done for dry and rainy season conditions at seven locations (S1, S2, S3, ..., S7) along the river, spaced 3 to 6 km apart to a total of 30.4 km. The results were grouped into four zones: (Z1) reference, (Z2) transition, (Z3) polluted, (Z4) recovery. The Alborada (S5) and Tunzingo (S6) sites, adjacent to dense high-class residential areas (Z3), had the greatest pollution charges in both seasons, while the La Poza (S7) site near the Tres Palos lagoon (Z4) showed a decrease in pollution. All parameters correlated with increasing head- to down-river sampling distance by following linear (pH, DO) or curvilinear patterns (all other parameters). Using sampling location and dry versus rainy sampling season as multivariate regression (predictor) variables led to least-squares capturing: 1) 66% to 95% of the T(°C), pH, DO, and PO_4^{3-} variations, and 2) 57% to 96% of the log-linear variations of the other parameters. Among the parameters,

T(°C), DO, and PO_4^{3-} were not significantly affected by sampling season, while pH became so after deleting two higher than usual pH values at the S5 and S6 locations during the dry season.

Keywords

Rio de la Sabana, Urban Developments, Water Pollution, Sampling Locations, Dry/Rainy Season, Multivariate Analyses

1. Introduction

Accelerated urbanization processes are causing water-quality disruptions within rivers and streams across watersheds and regions. This is in part due to insufficient city planning towards environmental sustainability [1] which can lead to serious deteriorations in watershed- and river-based environmental and ecological services. Apart from noticing enhanced levels of water pollution and reduced biological viability in natural waterways, surficial water quality evaluations are needed to provide quantitative information for sustainable water-resource management [2] [3] [4]. Since these evaluations need to address a host of sampled water quality indicators, it is necessary to determine how these indicators are patterned and vary across space and time in a predictive manner as advocated by, e.g., [5]-[10]. The results so obtained can then be transformed into useful decision-supporting tools to restore and protect water quality locally to regionally [3] [9] [11]. This article focuses on assessing water quality changes along a 30.4 km stretch of Rio de la Sabana as it flows east of Acapulco into the coastal Tres Palos lagoon (Figure 1). During the last two decades, the mid-low part of this sub-basin has experienced an accelerated population growth, which has led:

- 1) to serious water pollution and environmental degradation issues towards the river and to the coastal lagoons at Tres Palos and Puerto Marques [12] [13];
- 2) to increased vulnerability to health risks across the new and flood-prone de la Sabana Valley settlements [13].

For these reasons, the Mexican government with support from the Spanish government initiated a potable water supply and wastewater sanitation project in 2012 to improve the quality of life, promote social equity and environmental sustainability across the de la Sabana Valley [12]. As such, the Rio de la Sabana River—Tres Palos Lagoon area is now part of the hydrological priority region for preserving the biological importance of the coastal and marine Mitla-Chautengo Lagoon System [14]. To understand and quantify how water quality varies along Riode la Sabana from its headwaters to the coastal lagoon, dry and rainy season water samples were taken from seven locations, all spaced 3 to 6.5 km apart. These locations vary from low to densely populated areas. The samples were analyzed to inform about 14 physicochemical and biological parameters (temperature, pH, electrical conductivity, dissolved oxygen, biochemical and

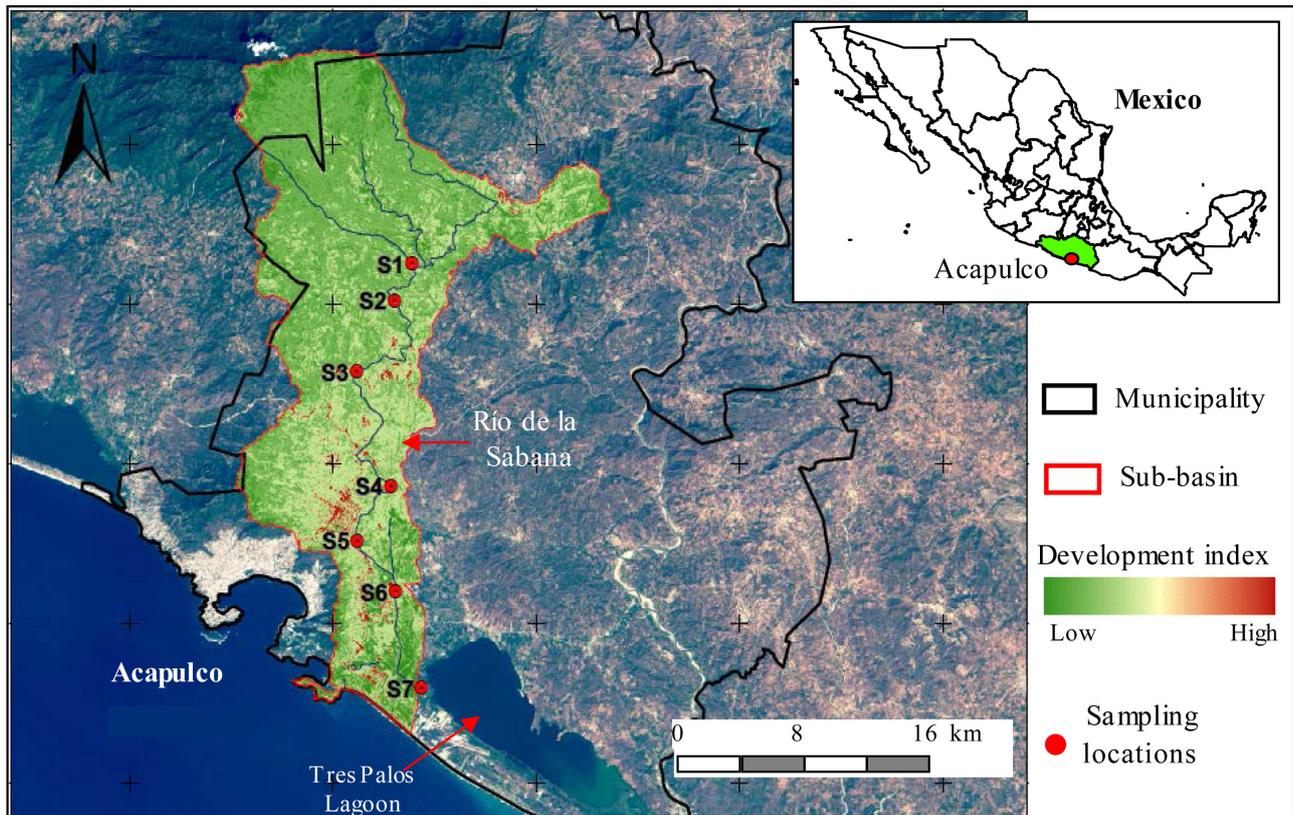


Figure 1. Locator map centered on the Rio La Sabana—Tres Palos Lagoon sub-basin north to east within the Acapulco Municipality in southwest Mexico also showing the seven sampling locations along the river, with a development index from low to high, inversely related to the normalized difference vegetation index (NDVI).

chemical oxygen demand, ammonium, nitrate, nitrite, sulphate, phosphate, methylene blue active substances, total coliforms, fecal coliforms).

2. Materials and Methods

Study area.

Rio de la Sabana begins north of Acapulco, with its headwater channels reaching up to about 2000 meters above sea level. Its main flow channel is approximately 57 kilometers long at its confluence with the Tres Palos coastal lagoon [12]. The sub-basin area amounts to 466.3 km² with combined permanent flow channel lengths of 727 km, thus leading to a drainage density of 1.55 km⁻¹ [15]. The upper part of the sub-basin is comprised of steep sparsely populated terrain with many short flow channels fed by short-duration runoff. In contrast, the lower part widens into a broad floodplain, of which the eastern part is now densely populated [12].

Sampling strategy.

Water quality sampling was conducted in 2017 during the dry season (January) and rainy season (July) in 7 sites (Table 1) from the headwater to the mouth of Rio de la Sabana at the Tres Palos lagoon (Figure 1). These locations were selected based on Google Earths images and mapped inland-use, vegetation cover,

Table 1. Location of study sites for water-quality characterization.

Reach	Sampling location	Location name	Latitude	Longitude	Elevation m
Upper	S1	km 39	17°2'47.28" N	99°46'57.24" W	400
	S2	km 34	17°1'27.84" N	99° 47'38. 41" W	300
	S3	Puente Texca	16°58'55.62" N	99°49'6.51" W	100
	S4	Colonia La Venta	16°55'25.26" N	99°48'27.28" W	40
Middle	S5	Puente Alborada	16°52'48.13" N	99°49'6.00" W	20
	S6	Puente Tunzingo	16°51'1.09" N	99°47'37.38" W	20
Lower	S7	Colonia La Poza	16°47'33.09" N	99°46'56.16" W	3

hydrology and population density [16]. Pre-sampling involved 12 sites located in areas with similar characteristics (land inclination, substrate type, water flow rate, and shading degree) to analyze 7 physicochemical parameters (temperature, pH, electrical conductivity, dissolved suspended solids, residual chlorine, iron and magnesium). The resulting spatial and physicochemical analyses enabled the selection of 3 sites on the upper scarcely populated section (reference zone), and 4 sites in the densely populated mid to low river sections (Figure 3). Within the impacted zone, there are numerous pollution sources due to solid waste dumpsites and industrial and household wastewater discharge [12] [13].

Parameter monitoring and analytical methods.

Fourteen physicochemical, biochemical and bacteriological water quality parameters (Table 2) were measured in water samples in each of the study sites, following established protocols [17]. Temperature, pH and electrical conductivity were measured *in situ*. The other parameters were analyzed at a laboratory accredited by the Mexican Organization of Accreditation. Temperature, pH, electrical conductivity, ammonium, nitrates, nitrites, methylene blue active substances, sulfates and phosphates parameters were all measured in triplicate. Dissolved oxygen, biochemical oxygen demand, chemical oxygen demand, and total and fecal coliforms were determined once using certified testing procedures. The results were organized in a data matrix (sites x parameters), where the rows refer to the sampling locations and the columns refer to the analyzed parameters (Table 2). These results were subsequently plotted by sampling location and sampling season. The order between the sample concentrations, locations, and season was examined by way of cluster analysis (CA). A correlation matrix was established to determine how the parameters relate to each other, to sampling location, and to season by way of principle component analysis. This was followed by nonlinear and linear least-squares multiple regression analyses to determine how each parameter varied quantitatively by sampling location and season. The non-linear analysis was based on the following equation:

$$y = a_{NL} + b_{NL} \exp\left[-\left(\frac{x - c_{NL}}{d_{NL}}\right)^2\right] \quad (1)$$

Table 2. Physicochemical and bacteriological parameters, units and methods used.

	Parameter	Unit		Analytical Method
1	Temperature	T	°C	Mercury thermometer
2	pH	pH	No unit	Potentiometric
3	Electrical Conductivity	EC	µs/cm	Potentiometric
4	Dissolved Oxygen	DO	mg/L	Sodium Azide (Winkler)
5	Biochemical Oxygen Demand	BOD ₅	mg/L	5-day Oxygen Consumption
6	Chemical Oxygen Demand	COD	mg/L	Potassium Dichromate Organic Matter Oxidation
7	Ammonium Nitrogen	NH ₃	mg/L	Kjeldahl Distillation
8	Nitrates	NO ₃ ⁻	mg/L	Brucine Sulfate (UV spectrometric)
9	Nitrites	NO ₂ ⁻	mg/L	Sulphanilamide
11	Sulfates	SO ₄ ²⁻	mg/L	Barium Chloride (UV spectrometric)
12	Phosphates	PO ₄ ³⁻	mg/L	Vanadomolybphosphoric Acid
10	Methylene Blue Active Substances	MBAS	mg/L	Methylene Blue
13	Total Coliforms	TC	NMP/mL	Multiple-Tube Fermentation Technique
14	Fecal Coliforms	FC	NMP/mL	Multiple-Tube Fermentation Technique

and the linear multiple regression analysis used the following equation:

$$y = a_L + b_L x + c_L \text{ season} \quad (2)$$

with y representing any of the 14 parameters, x representing the sampling locations S1, S2, S3, ..., S7, a_L and a_{NL} refer to the intercepts, and b_L , c_L , b_{NL} , c_{NL} , d_{NL} are regression coefficients. The resulting best-fitted extent of the parameter variations, indicated by the coefficient of variation (R^2) and the root mean square of the residuals, was improved for some of the parameters through log transformation. Equation (1) was applied to generate the best-fitted lines for each parameter by season, with sampling locations coded 1, 2, 3, 4, 5, 6, 7 according to their original order. Equation (2) was used to test the statistical significance of each parameter by location and season.

3. Results and Discussion

The data for the 14 water quality parameters are listed in **Table 3** and are plotted in **Figure 2** by sampling location and season, with best-fitted Equation (1) results overlaid. As tabulated and as shown, most parameters except for pH and DO increased from S1 (representing the least populated area) towards the populated areas represented by S5 and S6, and dropped again at S7. In addition, most parameters increased from the rainy to the dry season due to lack of dilution, except 1) DO decreased slightly (due to higher flow rates and better aeration), but not significantly so likely due to low sampling density, and 2) water temperature remained more or less the same. The parameters that increased the most at S5

Table 3. Sampling results, by sampling location and season.

Season	Location	Description	T (°C)	pH	EC	BOD ₅	COD	DO	NH ₃	NO ₃ ⁻	NO ₂ ⁻	SO ₄ ²⁻	PO ₄ ³⁻	MBAS	TC	FC
Dry	S1	km 39	22.6	8.44	160.5	39.2	60.2	5.42	0.2	9.2	0.08	50.2	0.02	0.15	9300	9300
	S2	km 34	26.9	8.23	190.6	39.3	65.5	5.84	0.2	20.1	0.14	49.8	0.04	0.14	9300	9300
	S3	Pte. Texca	27.7	8.26	329.0	47.1	76.0	5.35	0.6	28.6	0.21	75.2	0.08	0.32	150,000	150,000
	S4	Col. La venta	34.9	8.86	498.0	60.3	100.5	5.05	9.5	30.0	4.12	95.2	0.11	10.1	210,000	210,000
	S5	Pte. Alborada	32.6	7.96	1020.0	156.4	284.3	0.89	10.3	32.0	17.3	195.6	0.15	22.4	210,000	210,000
	S6	Pte. Tunzingo	35.3	9.40	1022.0	165.4	295.4	1.80	11.1	36.1	11.1	215.4	0.16	23.8	930,000	930,000
	S7	Col. La poza	33.4	7.67	887.0	137.3	280.2	2.05	9.4	28.6	6.94	208.3	0.14	18.6	150,000	150,000
Rainy	S1	km 39	24.0	7.51	73.7	5.3	11.2	6.23	0.26	4.27	0.01	15.28	0.01	0.10	1	1
	S2	km 34	26.9	7.69	94.4	7.5	15.3	6.15	0.30	4.54	0.01	14.52	0.01	0.10	2.7	2.0
	S3	Pte. Texca	28.2	7.49	123.8	10.5	20.4	5.87	0.31	6.63	0.02	21.04	0.02	0.15	26.7	15.0
	S4	Col. La venta	32.8	7.29	201.0	13.7	31.7	5.27	0.57	7.89	0.16	22.42	0.14	0.65	13.7	6.0
	S5	Pte. Alborada	32.6	7.23	569.1	36.8	92.7	1.86	5.13	25.24	0.27	90.00	0.19	9.25	79.0	68.7
	S6	Pte. Tunzingo	32.9	7.16	451.1	31.7	73.4	2.65	2.95	17.58	0.29	37.83	0.16	7.96	70.7	56.7
	S7	Col. La poza	33.1	7.24	473.9	33.5	79.7	2.38	3.22	19.01	0.15	78.27	0.15	7.90	24.7	14.0

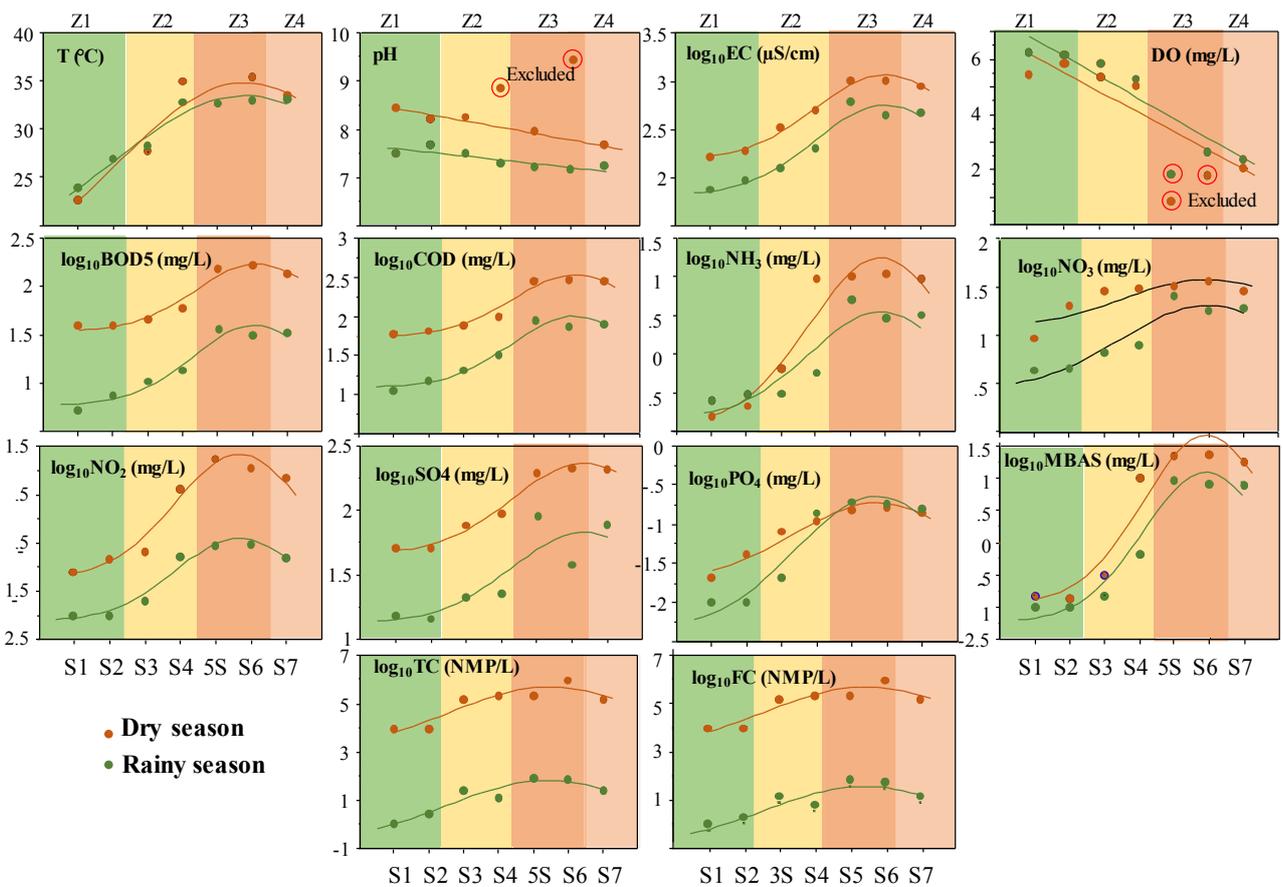


Figure 2. Scatter plots and best-fitted lines (Equation (2)) for each of the 14 parameters listed in Table 3, by sampling location (S1, S2, ... S7 along x-axis) and by pollution zone, colour-differentiated for each parameter from left to right by pollution level: nearly pristine (Z1), transitional (Z2), most polluted (Z3), and somewhat diluted (Z4).

and S6 from the rainy to the dry season were EC, BOD₅, COD, NO₃⁻, NO₂⁻, TC and FC. There were two notable outliers for pH during the dry sampling season at S4 (pH = 8.86) and S6 (pH = 9.4), somewhat parallel to high effluent concentrations pertaining to NH₃, NO₃⁻, MBAS, TC and TF. Also notable is the steep decline of DO at S5 and S6 for both the dry and rainy sampling season. This undoubtedly relates to elevated BOD and COD discharge at these locations, as also reported in [12] [13].

The best-fitted Equation (1) results for the parameters with non-linear S1 to S7 trends are listed in **Table 4**. The similarities among these trends are such that the c_{NL} and d_{NL} coefficients could be kept in common without significant R² and RMSE differences. In detail:

- 1) a_{NL} refers to the headwater values for each water parameter by season;
- 2) b_{NL} quantifies the pollution extent for each water parameter by season;
- 3) $c_{NL} = 5.58$ indicates that the maximum levels are associated with the S5 and S6 locations;
- 4) $d_{NL} = 2.60$ quantifies the spatial pollution extent across the S1 to S7 sampling locations.

The results in **Table 4** are used to estimate maximum and minimum water pollution levels along the river transect for the dry and rainy season (**Table 5**). In turn, the ratios of these numbers can be interpreted as spatial and seasonal pollution indicators. For example, transitions from rainy to dry season led to a min/min and max/max fecal count multiplication factors of about 7600 at the headwater region and 11,500 at the maximum pollution locations. By rainy to dry season, the down-river fecal pollution count increased by a max/min

Table 4. Best-fitted non-linear regression analysis results (Equation (1)); pH and DO not included.

Variable	Dry season				Rainy season				R ²	RMSE
	a_{NL}		b_{NL}		a_{NL}		b_{NL}			
	coeff.	st. error	coeff.	st. error	coeff.	st. error	coeff.	st. error		
T(°C)	23.20	1.40	13.40	2.00	24.40	1.30	9.70	1.90	0.885	2.81
log ₁₀ EC	2.18	0.05	0.88	0.07	1.82	0.04	0.93	0.07	0.979	0.004
log ₁₀ BOD ₅	1.49	0.06	0.72	0.10	0.71	0.06	0.86	0.07	0.971	0.0083
log ₁₀ COD	1.68	0.07	0.80	0.11	1.01	0.07	0.85	0.11	0.847	0.01
log ₁₀ NH ₃	-0.93	0.17	2.20	0.25	-0.82	0.17	1.38	0.25	0.915	0.057
log ₁₀ NO ₃ ⁻	1.14	0.09	0.49	0.15	0.45	0.12	0.85	0.16	0.902	0.015
log ₁₀ NO ₂ ⁻	-1.23	0.02	2.55	0.02	-2.13	0.17	1.71	0.21	0.980	0.027
log ₁₀ MBAS	-0.95	0.26	2.60	0.31	-1.23	0.24	2.29	0.30	0.950	0.079
log ₁₀ SO ₄ ²⁻	1.66	0.10	0.71	0.15	1.12	0.10	0.70	0.15	0.902	0.21
log ₁₀ PO ₄ ³⁻	-1.61	0.11	0.89	0.16	-2.10	0.12	1.58	0.17	0.937	0.0207
log ₁₀ FC	4.02	0.25	1.74	0.37	0.10	0.25	1.60	0.37	0.980	0.124
log ₁₀ TC	4.02	0.12	1.74	0.37	0.24	0.12	1.67	0.37	0.982	0.0987

$c_{NL} = 5.58 \pm 0.23$; $d_{NL} = 2.60 \pm 0.49$.

Table 5. Determining the minimum and maximum pollution levels and the corresponding ratios for the 12 parameters listed in Table 4, by location and by season (pH and DO not included).

Parameter	Dry Season		Rainy Season		Dry/Rainy Season		Location max/min	
	min ^a	max ^b	min ^a	max ^b	min/min	max/max	dry	rainy
T(°C)	23.2	36.6	24.4	34.1	0.95	1.1	1.6	1.4
EC	151.4	1148.2	66.1	562.3	2.3	2.0	7.6	8.5
BOD ₅	30.9	162.2	5.1	37.2	6.0	4.4	5.2	7.2
COD	47.9	302.0	10.2	72.4	4.7	4.2	6.3	7.1
NH ₃	0.12	18.62	0.15	3.63	0.78	5.1	158	24.0
NO ₃ ⁻	13.8	42.7	2.8	20.0	4.9	2.1	3.1	7.1
NO ₂ ⁻	0.059	20.9	0.007	0.380	7.9	55.0	355	51.3
SO ₄ ²⁻	45.7	234.4	13.2	66.1	3.5	3.5	5.1	5.0
PO ₄ ³⁻	0.025	0.191	0.008	0.302	3.1	0.6	7.8	38.0
MBAS	0.11	44.7	0.06	11.5	1.9	3.9	398	195
FC	6,456	588,843	0.9	51.3	7586	11482	91.2	60.3
TC	6,456	354,813	1.7	81.3	3715	4365	55.0	46.8

^amin: a_{NL} for T(°C) and 10^{aNL} for all other parameters. ^bmax: a_{NL} + b_{NL} for T(°C) and 10^{aNL+bNL} for all other parameters.

multiplication factor of 60 and 90, respectively. For the other parameters, down-river changes in pollution were most severe for NH₃, NO₃⁻, NO₂⁻, PO₄³⁻ and MBAS, with max/min pollution effects stronger for NO₃⁻ and PO₄³⁻ during the rainy season, and stronger for MBAS, NH₃ and NO₂⁻ during the dry season.

The seasonal NO₃⁻ to NO₂⁻ and NH₃ differences were likely related to the increasing extent to which added NO₃⁻ is converted to NO₂⁻ and NH₃ as the river flow rate drops from the rainy to the dry season. In this regard, it can be determined from Table 3 that:

- 1) the total N concentrations within the river water (NO₃⁻, NO₂⁻, NH₃ combined) decreased on average from the dry to the rainy season by a factor of 3.5, likely due to dilution;
- 2) at S4, S5, S6, and S7, the combined NO₂⁻ and NH₃ concentrations amounted to 60% of total N during the dry season, and about 40% during the wet season (mostly NH₃ only);
- 3) the least amount of water-dissolved N in the form of NO₂⁻ and NH₃ occurred at S1, S2, and S3 during the dry season (<10%), and increased to about 20% during the wet season.

Cluster analysis. The CA-generated dendrograms are displayed by dry and rainy sampling season in Figure 3 also presenting the normalized parameter values in Table 2. These dendrograms show that locations S1, S2, S3 and S4 are grouped within Cluster A, thereby representing the upper part of the Rio de la Sabana watershed where industrial to residential discharge rates into the river are still low. Within this cluster, pollution levels increased in the order

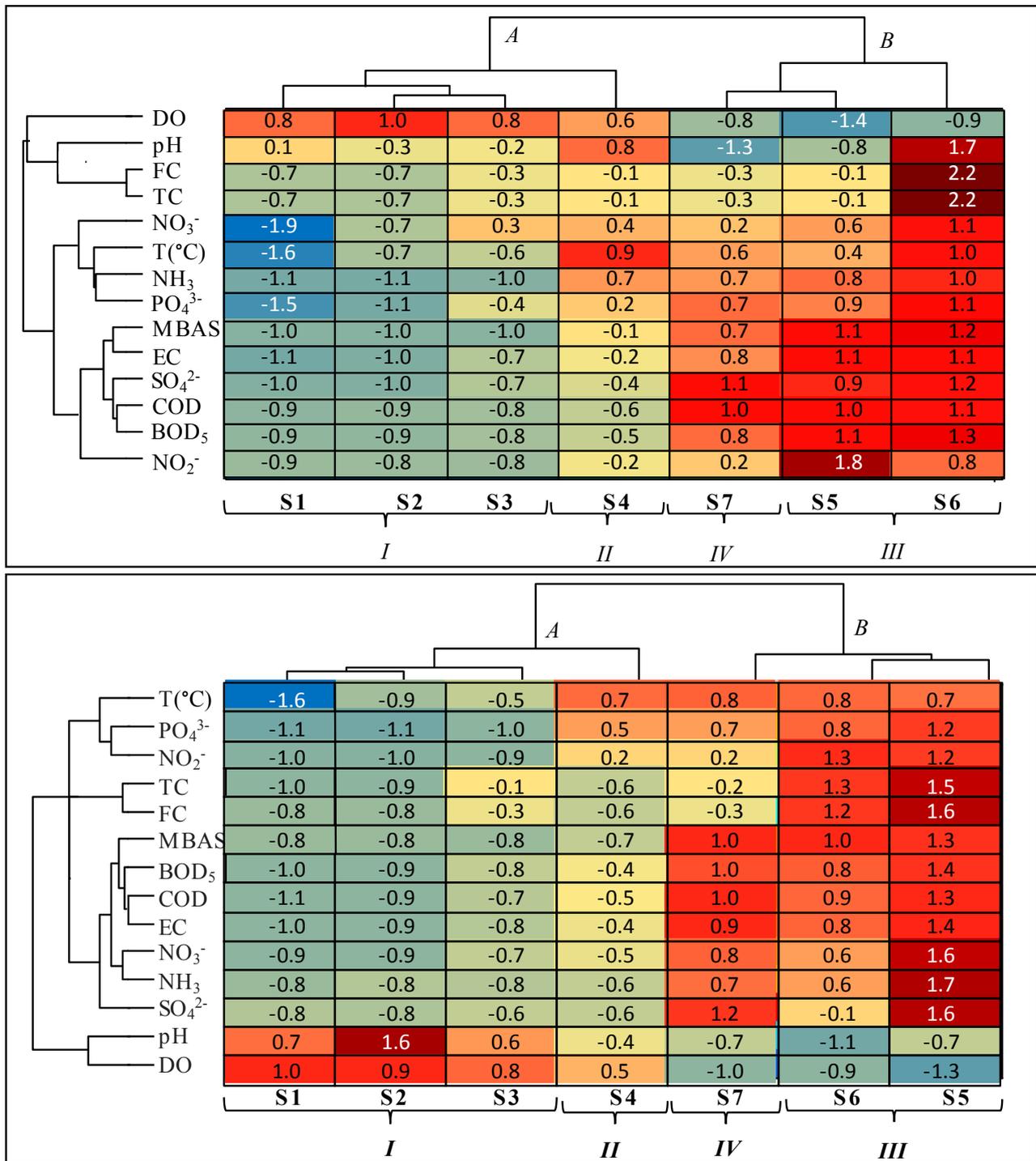


Figure 3. Zoning the physicochemical and bacteriological sampling results of Rio de la Sabana by way of hierarchical cluster dendrograms. Also shown: normalized parameter scores overlaid on value-coded background (from blue to yellow to dark red) by sampling season Top: dry season; bottom: rainy season. Note the seasonal reversal of the S5 and S6 locations.

S1 < S2 < S3 < S4, *i.e.*, in direct relation to increased settlement densities. The densely inhabited areas at S5 (Puente Alborada), S6 (Puente Tunzingo) and S7 (Colonia La Poza) at the mid to low portions of the Rio de la Sabana sub-basin form Cluster B, owing to high industrial to residential discharge rates [12] [13].

During the dry season, the increases from S1 to S2 and S3 remained similar to each other, but the increases for S4 (located in an area with a greater presence of rural towns) were more similar to the elevated results for the Cluster B locations. Within Cluster B, the dendrogram positions switched from $S7 < S5 < S6$ during the dry to $S7 < S6 < S5$ during the rainy season. This suggests that overall water quality:

1) recovered somewhat for both seasons after passing through the S5 and S6 locations, presumably due to the influx of less contaminated floodplain water from the eastern less inhabited part of the Rio de la Sabana watershed;

2) was worst at S5 during the rainy season where pollutant inputs are likely highest due industrial and residential surface run-off;

3) was worst at S6 during the dry season mainly due to accumulating up-river sewage discharge during low river flow rates.

Based on location similarities, Clusters A and B were divided into four zones:

1) The Reference zone (S1, S2): located at the sub-basin's upper part, meeting national and international standards established for aquatic life protection [18]-[22].

2) The Transition zone (S3, S4): a peri-urban zone located between the sub-basin's upper and mid-low parts had all parameters except pH and DO rising from S3 to S4 above their values at S1 and S2.

3) The Pollution zone (S5, S6): all parameters except pH and DO were well above their average values and above national and international standards for aquatic life protection [18]-[22].

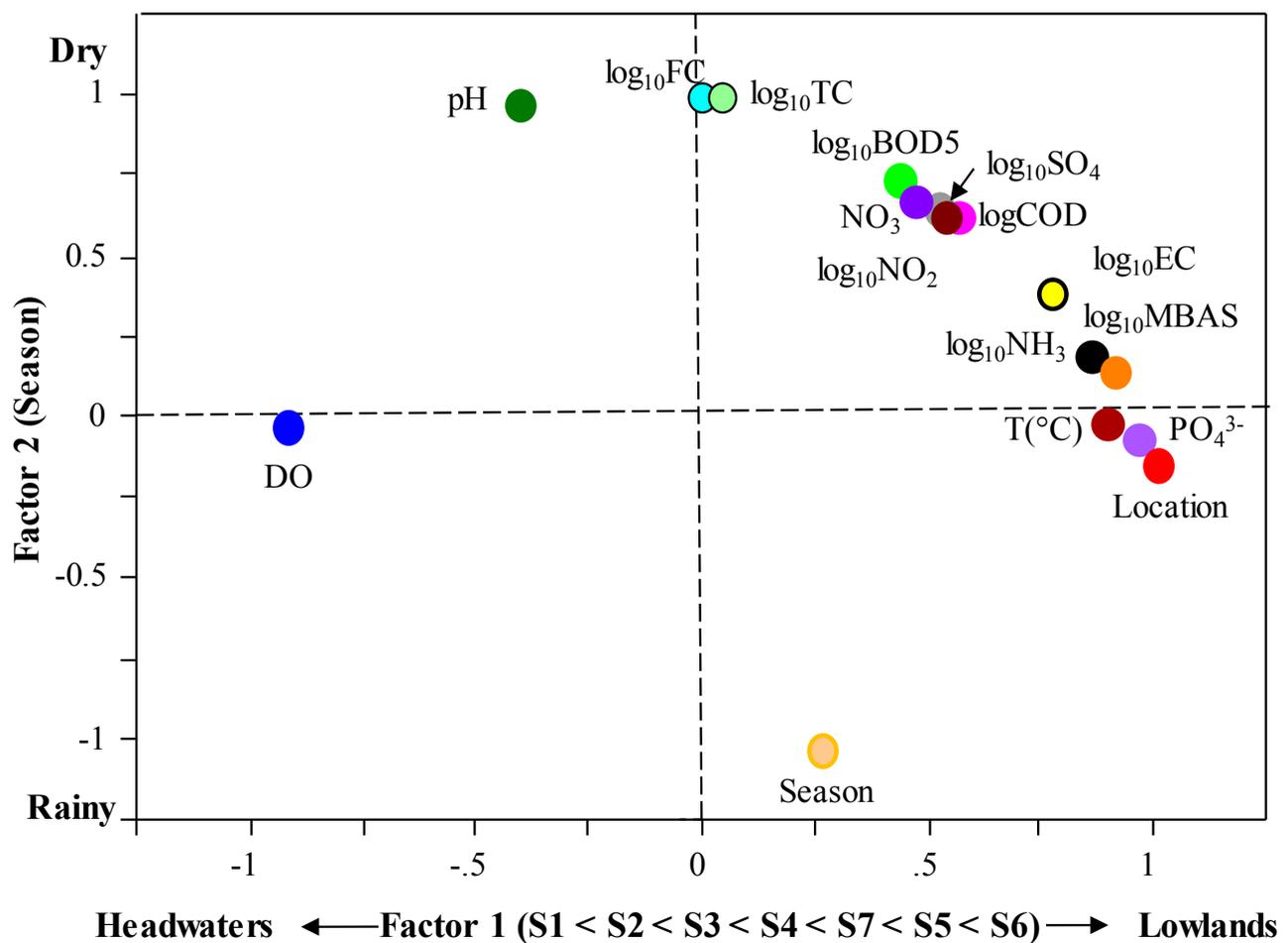
4) The Recovery zone (S7).

Correlation matrix. All parameters other than DO and pH were highly positively correlated to one another during the rainy and the dry season, as shown in **Table 6**. Excluding the high pH values at S4 (pH = 8.86) and S7 (pH = 9.40) yielded a positive correlation between pH and DO for both season, with both remaining negatively related to all the other parameters. The generally gradual pH decline from S1 to S7 may be due to a greater presence of dissolved organic acids in response to a transition from upper-reach groundwater seepage to lower-reach surface run-off. The concurrent decrease in DO would be due to a stimulated chemical and biological oxygen demand due to enhanced oxygen-consuming fecal-matter containing effluents arriving at the S4 - S7 sampling locations [7] [12] [23] [24]. This would also include the discharge of surfactants (represented by MBAS) and detergents [17] (NMX-AA-039-SCFI-2001). Consequently, effluent-associated parameters such as EC, BOD₅, COD, NH₃, NO₂⁻, MBAS, SO₄²⁻ and PO₄³⁻ were all highly correlated to one another for both seasons [13] [25]. Their corresponding dry-to-rainy season reductions as plotted in **Figure 2**, are undoubtedly due to increased water flow dilution.

Analyzing the parameter correlation matrix with sampling location and season as two additional variables produced the Factor 2 versus Factor 1 plot in **Figure 4** by way of principle component analysis (PCA). Here, Factor 1 refers to the S1, S2, S3, S4, S7, S5, S6 sampling sequence, while Factor 2 refers to the rainy (coded 0) versus the dry (coded 1) season. In **Figure 4**, the parameters entering

Table 6. Correlation matrix for the 14 parameters in **Table 2**.

	T(°C)	pH	log ₁₀ EC	DO	log ₁₀ BOD ₅	log ₁₀ COD	log ₁₀ NH ₃	NO ₃ ⁻	log ₁₀ NO ₂ ⁻	log ₁₀ SO ₄ ²⁻	PO ₄ ³⁻	log ₁₀ MBAS	log ₁₀ TC	log ₁₀ FC
T(°C)	1.000	0.270	0.891	-0.659	0.783	0.783	0.959	0.900	0.922	0.833	0.918	0.927	0.866	0.866
pH	-0.856	1.000	0.058	0.088	0.035	-0.018	0.150	0.237	0.103	0.015	0.122	0.133	0.371	0.371
log ₁₀ EC	0.906	-0.891	1.000	-0.910	0.962	0.959	0.954	0.869	0.970	0.986	0.996	0.963	0.884	0.884
DO	-0.800	0.844	-0.978	1.000	-0.977	-0.973	-0.785	-0.641	-0.877	-0.944	-0.876	-0.851	-0.671	-0.671
log ₁₀ BOD ₅	0.910	-0.872	0.996	-0.968	1.000	0.997	0.867	0.744	0.929	0.986	0.943	0.917	0.776	0.776
log ₁₀ COD	0.909	-0.884	0.999	-0.975	0.998	1.000	0.863	0.737	0.924	0.986	0.938	0.913	0.754	0.754
log ₁₀ NH ₃	0.816	-0.848	0.981	-0.998	0.970	0.978	1.000	0.850	0.977	0.913	0.958	0.988	0.876	0.876
NO ₃ ⁻	0.760	-0.801	0.958	-0.985	0.947	0.954	0.983	1.000	0.819	0.799	0.908	0.796	0.930	0.930
log ₁₀ NO ₂ ⁻	0.952	-0.954	0.939	-0.867	0.924	0.934	0.883	0.837	1.000	0.942	0.965	0.991	0.823	0.823
log ₁₀ SO ₄ ²⁻	0.765	-0.780	0.940	-0.959	0.931	0.938	0.951	0.969	0.802	1.000	0.972	0.941	0.837	0.837
PO ₄ ³⁻	0.928	-0.928	0.949	-0.897	0.925	0.942	0.915	0.879	0.986	0.850	1.000	0.957	0.915	0.915
log ₁₀ MBAS	0.869	-0.909	0.990	-0.985	0.979	0.987	0.988	0.952	0.930	0.927	0.944	1.000	0.836	0.836
log ₁₀ TC	0.848	-0.779	0.863	-0.789	0.894	0.873	0.788	0.797	0.837	0.760	0.784	0.806	1.000	1.000
log ₁₀ FC	0.790	-0.775	0.873	-0.831	0.899	0.881	0.828	0.847	0.819	0.784	0.779	0.829	0.984	1.000

**Figure 4.** Scatter plot of Factor 2, mostly associated with dry versus rainy season sampling, versus Factor 1, mostly associated with sampling location sequenced from left to right as follows: S1, S2, S3, S4, S7, S5, S6.

towards the right are most positively related to sampling location while the parameters entering towards the top are most positively related to season (Factor 2). Both DO and pH were negatively influenced by sampling location (Factor 1), with pH positively related to the transition from the rainy to the dry season, while DO was not much influenced by this transition. The PO_4^{3-} and $\text{T}(\text{°C})$ parameters are shown to be closely and positively related to sampling location but not to sampling season.

Multiple regression analysis. The best-fitted and most significant intercepts (a_L) and regression coefficients (b_L , c_L) for Equation (2) are listed in **Table 7** for all 14 parameters ($p < 0.1$). Also entered are the corresponding R^2 and RMSE values, to indicate the extent of variation capture and associated regression error for each parameter. As shown, all parameters were significantly related to sampling location. Only three parameters, namely $\text{T}(\text{°C})$, PO_4^{3-} and DO, were not significantly affected by season. With DO, the increase from the dry to the rainy season was not significant at $p < 0.1$, even after excluding the low DO results at S4 and S5 (dry season) and S5 (rainy season). With pH, season sampling became significant after excluding the high dry-season pH values at the S4 and S6 (pH = 8.86 and 9.4) from the analysis.

Altogether, the results in **Table 7** confirm that the location-specific and season-specific effects on water pollution are highly significant and can now in part

Table 7. Best-fitted Equation (2) regression results for each parameter listed in **Table 3**, with sampling location, season and Zone 4 (S7) as independent regression variables.

Parameter	Intercept: a_L		Regression coefficients				R^2	RMSE
			Location: b_L^a		Season: c_L^b			
	Est.	St. Error	Est.	St. Error	Est.	St. Error		
$\text{T}(\text{°C})$	23.4	1.6	1.7	0.3	NS		0.865	1.73
pH	7.6	0.3	-0.62	0.06	NS		0.663	0.43
pH^c	7.7	0.1	-0.87	0.02	-0.87	0.08	0.923	0.14
log₁₀EC	1.70	0.06	0.16	0.01	0.34	0.05	0.648	0.09
DO	7.8	0.5	-0.87	0.11	NS		0.861	0.80
log₁₀BOD₅	0.63	0.07	0.14	0.01	0.69	0.06	0.570	0.11
log₁₀COD	0.92	0.07	0.15	0.02	0.58	0.06	0.948	0.11
log₁₀NH₃	-1.2	0.2	0.30	0.04	0.36	0.18	0.824	0.33
log₁₀NO₃⁻	0.6	0.1	1.10	0.02	0.40	0.07	0.859	0.13
log₁₀NO₂⁻	-2.6	0.2	0.36	0.05	1.4	0.2	0.901	0.37
log₁₀SO₄²⁻	0.96	0.07	0.13	0.01	0.54	0.06	0.946	0.10
log₁₀PO₄³⁻	-2.1	0.2	0.20	0.03	NS		0.783	0.24
log₁₀MBAS	-1.8	0.4	0.43	0.05	0.44	0.21	0.862	0.40
log₁₀TC	0.1	0.3	0.27	0.06	3.8	0.2	0.962	0.00
log₁₀FC	0.1	0.3	0.27	0.06	4.0	0.2	0.965	0.44

^aLocation: S1, S2, S3, S4, S7, S5, S6 coded 1, 2, 3, 4, 5, 6, 7 (note the change in order for S5, S6 and S7).

^bSeasons coded 0 (rainy), 1 (dry); NS: not significant. ^cpH outliers excluded.

be interpolated for any location along Rio de la Sabana. This can be done by way of the best-fitted Equation (1) and Equation (2) regression results, and by ranking specific locations using existing conditions at S1, S2, ..., S7 as a guide. To this effect, locations along the Tres Palos lagoon may eventually experience pollution levels similar to the S5 and S6 locations. However, improved pollution mitigation may lower the pollutant levels at S4 to S7 locations through, e.g., biological denitrification, sulphate and phosphate removal by way of chemical and biological means, and effluent sterilization.

While DO is negatively related to all the other water quality parameters as well as location, its overall variations are best captured by way of the following multivariate regression equation, and as plotted in **Figure 5**:

$$\text{DO} = (0.84 \pm 0.03) \text{pH} - (0.0059 \pm 0.0005) \text{EC}; R^2 = 0.923; \text{RMSE} = 0.59. \quad (3)$$

In principle, this equation reflects the association of higher DO and pH and lower EC values in the upper river reach compared to the lower river reach. The lack of an inverse relationship between DO, BOD5 and COD is likely due to other DO influencing factors such as the diurnal oxygen release from photosynthesizing plants and algae within the river [26].

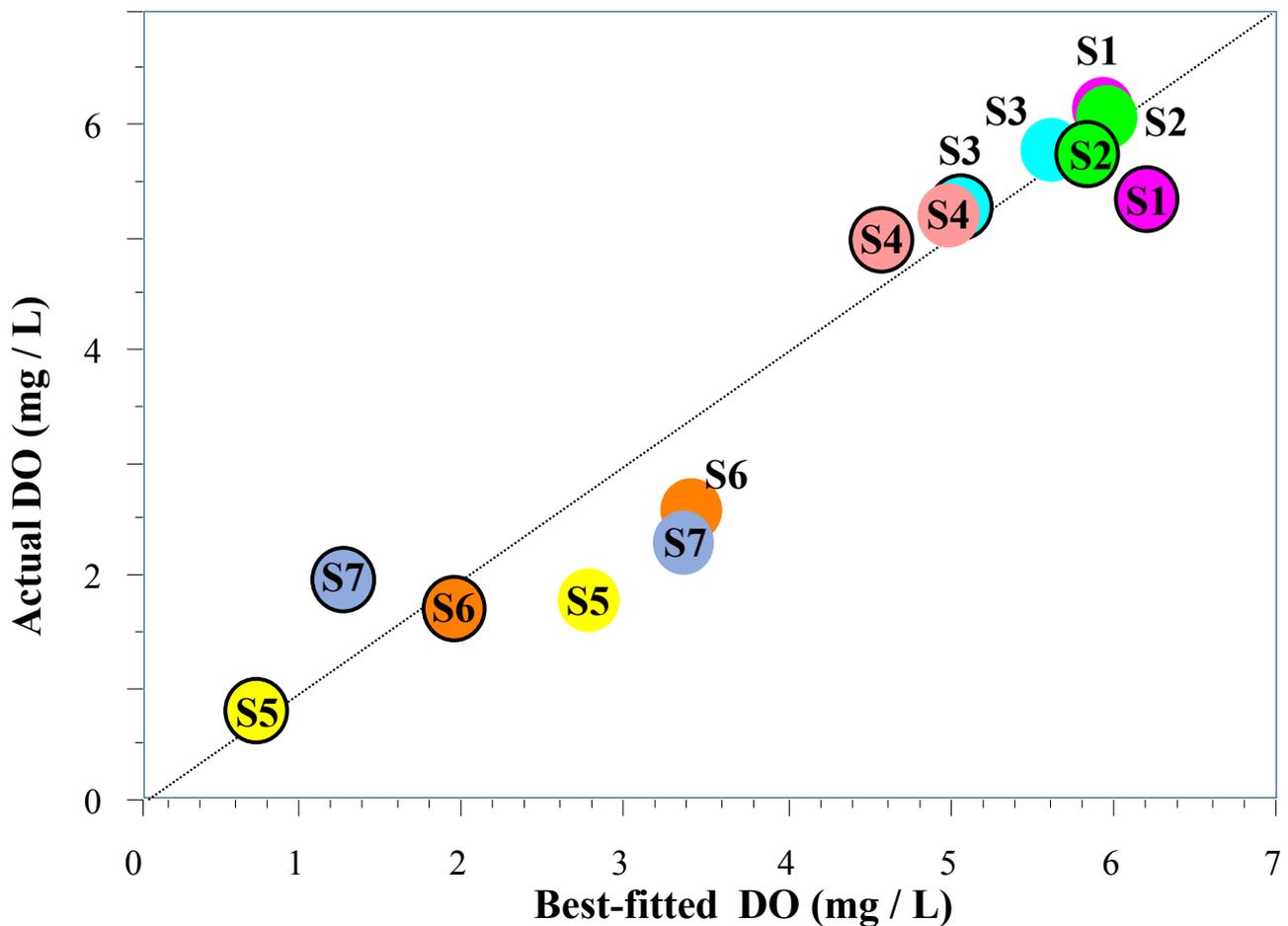


Figure 5. Plotting actual versus best-fitted S1 to S7 DO values generated with Equation (3). Differences by season are marked by dot outline: none for rainy season; black for dry season.

4. Conclusions

During the last decades, increased developments on the east and west sides of the Rio de la Sabana floodplain have led to extensive water pollution, with higher pollution levels registered for the dry and rainy seasons. This study marks the S5 and S6 locations as currently the most polluted locations. Upriver, the S4 location may also become increasingly vulnerable to pollution. Downriver, there was a slight reduction in water pollution, likely due to dilution caused by water seepage and run-off from the yet fairly undeveloped low-lying areas on the east and northern side of the river basin.

The analysis of this sampling effort revealed that all 14 water quality parameters were significantly related to sampling locations where current settlement densities and consequently effluent discharge rates would be highest. By induction, the above approach could prove useful: 1) in application at other settlement-affected locations, especially those along stream and effluent discharges towards the coastal Tres Palos and Puerto Marques lagoons, and 2) in encouraging initiatives towards intensifying wastewater treatment along Rio de la Sabana developments and elsewhere.

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Conflict of Interests

The authors have declared that no conflict of interests exists.

References

- [1] Sugandi, A., Astuti, W. and Erianto, E. (2017) Waterfront Development Concepts in Indonesia from the Perspective of Urban Planning and Environmental Sustainability. *International Journal of Built Environment and Sustainability*, **4**, 146-155.
- [2] Gholami, S. and Srikantaswamy, S. (2009) Statistical Multivariate Analysis in the Assessment of River Water Quality in the Vicinity of KRS Dam, Karnataka, India. *Natural Resources Research*, **18**, 235-247.
<https://doi.org/10.1007/s11053-009-9096-y>
- [3] Mohamed, I., Othman, F., Ibrahim, A., Alaa-Eldin, M. and Yunus, R. (2016) Assessment of Water Quality Parameters Using Multivariate Analysis for Klang River Basin, Malaysia. *Environmental Monitoring and Assessment*, **187**, 4182.
- [4] Martínez-Tavera, E., Rodríguez-Espinosa, P.F., Shruti, V.C., Sujitha, S.B., Morales-García, S.S. and Muñoz-Sevilla, S.P. (2017) Monitoring the Seasonal Dynamics of Physicochemical Parameters from Atoyac River Basin (Puebla), Central Mexico: Multivariate Approach. *Environmental Earth Sciences*, **76**, 95.
<https://doi.org/10.1007/s12665-017-6406-2>

- [5] Achieng', A., Raburu P., Kipkorir, E., Ngodhe, S., Obiero, K. and Ani-Sabwa, J. (2017) Assessment of Water Quality Using Multivariate Techniques in River Sosiani, Kenya. *Environmental Monitoring and Assessment*, **189**, 280.
- [6] Aruga, R., Negro, G. and Ostacoli, G. (1993) Multivariate Data Analysis Applied to the Investigation of River Pollution. *Fresenius Journal of Analytical Chemistry*, **346**, 968-975. <https://doi.org/10.1007/BF00322761>
- [7] Shrestha, S. and Kazama, F. (2007) Assessment of Surface Water Quality Using Multivariate Statistical Techniques: A Case Study of the Fuji River Basin, Japan. *Environmental Modelling and Software*, **22**, 464-475. <https://doi.org/10.1016/j.envsoft.2006.02.001>
- [8] Gutiérrez-Cacciabue, D., Teich, I., Poma, H., Cruz, M., Balzarini, M. and Rajal, B. (2014) Strategies to Optimize Monitoring Schemes of Recreational Waters from Salta, Argentina: A Multivariate Approach. *Environmental Monitoring and Assessment*, **186**, 8359-8380. <https://doi.org/10.1007/s10661-014-4010-4>
- [9] Sharma, M., Kansal, A., Jain, S. and Sharma, P. (2015) Application of Multivariate Statistical Techniques in Determining the Spatial Temporal Water Quality Variation of Ganga and Yamuna Rivers Present in Uttarakhand State, India. *Water Quality, Exposure, and Health*, **7**, 567-581.
- [10] Zheng, L.-Y., Yu, H.-B. and Wang, Q.-S. (2016) Application of Multivariate Statistical Techniques in Assessment of Surface Water Quality in Second Songhua River Basin, China. *Journal of Central South University*, **23**, 1040-1051. <https://doi.org/10.1007/s11771-016-0353-z>
- [11] Wang, Y., Wang, P., Bai, Y., Tian, Z., Li, J., Shao, X., Mustavich, L. and Li, B. (2013) Assessment of Surface Water Quality via Multivariate Statistical Techniques: A Case Study of the Songhua River Harbin Region, China. *Journal of Hydro-Environment Research*, **7**, 30-40. <https://doi.org/10.1016/j.jher.2012.10.003>
- [12] CONAGUA, CAPASEG, CAPAMA. (2012) Plan operativo general: Proyecto de suministro de agua potable y saneamiento de las zonas marginadas del valle de la Sabana en el estado de Guerrero. Acapulco, Mexico: Comisión de Agua Potable, Alcantarillado y Saneamiento del Estado de Guerrero. http://www.fondodelagua.aecid.es/galerias/fcas/descargas/proyectos/pog/POG_MEX-001-B.pdf
- [13] Rodríguez-Herrera, A.L., Olivier-Salomé, B., López-Velasco, R., Barragán-Mendoza, M., Cañedo-Villareal, R. and Valera-Pérez, M.A. (2013) Contaminación y riesgo sanitario en zonas urbanas de la subcuenca del Río de la Sabana, ciudad de Acapulco, Mexico. *Revista Gestión y Ambiente*, **16**, 85-96.
- [14] Arriaga-Cabrera, L., Aguilar-Sierra, J., Alcocer-Durand, R., Jiménez-Rosenberg, E., Muñoz-López, E. and Vázquez-Domínguez, E. (2008) Regiones hidrológicas prioritarias. Escala de Trabajo 1:4 000 000. 2ª. Edición. Comisión Nacional para el Conocimiento y Uso de la Biodiversidad (CONABIO), Mexico. <http://www.conabio.gob.mx/conocimiento/regionalizacion/doctos/Hmapa.html>
- [15] Villegas-Romero, I., Oropesa-Mota, J.L., Martínez-Ménes, M. and Mejía-Sáenz, E. (2009) Trayectoria y relación lluvia-escurrimiento causados por el Huracán Paulina en la cuenca del Río de la Sabana, Guerrero, Mexico. *Agrociencia*, **43**, 345-356.
- [16] INEGI (2010) Instituto Nacional de Estadística y Geografía. Recursos naturales. <http://www.inegi.org.mx/geo/contenidos/reclnat/default.aspx>
- [17] NMX-AA-039-SCFI-2001 Waters Analysis—Determination of Methylene Blue Active Substances in Natural, Drinking, Wastewaters and Wastewaters Treated—Test Method. <http://legismex.mty.itesm.mx/normas/aa/aa039-01.pdf>

- [18] CE-CCA-001/89 (1989) Acuerdo por el que se establecen los criterios ecológicos de la calidad del agua. Centro De Calidad Ambiental (Uninet).
<http://legismex.mty.itesm.mx/acu/acca001.pdf>
- [19] CEQGs (1999) Canadian Water Quality Guidelines for the Protection of Aquatic Life Freshwater. Canadian Environmental Quality Guidelines, Canadian Council of Ministers of the Environment, Winnipeg.
- [20] CEQGs (2012) Canadian Water Quality Guidelines for the Protection of Aquatic Life: Nitrate. Canadian Environmental Quality Guidelines, Canadian Council of Ministers of the Environment, Winnipeg.
- [21] CONAGUA, Comisión Nacional del Agua (2015) Monitoreo de calidad del agua. Subdirección General Técnica. Gerencia de Calidad del agua.
<http://files.conagua.gob.mx/transparencia/CalidaddelAgua.pdf>
- [22] EPA (2017) National Recommended Water Quality Criteria-Aquatic Life Criteria Table. Water Quality Criteria.
<https://www.epa.gov/wqc/national-recommended-water-quality-criteria-aquatic-life-criteria-table>
- [23] Singh, K., Malik, A., Mohan, D. and Sinha, S. (2004) Multivariate Statistical Techniques for the Evaluation of Spatial and Temporal Variations in Water Quality of Gomti River (India)—A Case Study. *Water Research*, **38**, 3980-3992.
<https://doi.org/10.1016/j.watres.2004.06.011>
- [24] Araoye, P. (2009) The Seasonal Variation of pH and Dissolved Oxygen (DO₂) Concentration in Asa Lake Ilorin, Nigeria. *International Journal of Physical Sciences*, **4**, 271-274.
- [25] Ohrel, R. and Register, K. (2006) Oxygen. Volunteer Estuary Monitoring: A Methods Manual (9-1). U.S. Environmental Protection Agency (EPA), Office of Wetlands, Oceans and Watersheds, The Ocean Conservancy.
- [26] Zang, C., Huang, S., Wu, M., Du, S., Scholz, M., Gao, F., Lin, C., Guo, Y. and Doug, Y. (2011) Comparison of Relationships between pH, Dissolved Oxygen and Chlorophyll a for Aquaculture and Non-Aquaculture Waters. *Water, Air, & Soil Pollution*, **219**, 157-174. <https://doi.org/10.1007/s11270-010-0695-3>