

Separation of Benzene, Toluene, Ethylbenzene and P-Xylene from Aqueous Solutions by Carbon Nanotubes/Polyvinylidene Fluoride Nanocomposite Membrane

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Abstract

Carbon nanotubes/polyvinylidene fluoride (PVDF) nanocomposite membranes (abbreviated as CPMs) were fabricated to study their physicochemical property and separation efficiency of organic pollutants such as benzene, toluene, ethylbenzene and methylbenzene (abbreviated as BTEX) from aqueous solutions. The rejection coefficients (R) of BTEX depend on the CNT content of CPM, pore size of membrane, molecule size of BTEX, permeation pressure (P), feed concentration ($C_{\rm f}$) and temperature. The CNT contents were 5%, 10% and 15% in CPM have been conducted. The 10% CNT content of CPM (10-CPM) has not only higher water flux but also the relatively higher R as comparing of 5% and 15% CNT of CPM. The R decreased with increasing P, C_{c} and temperature but has no significant influence on ionic strength (μ). The R of BTEX were found in the order as B < T < E \approx X which revealed the major mechanism of BTEX separation with CPM was related to molecule size of BTEX (B < T < E \approx X). It exhibits that the size exclusion plays the important role in BTEX separation. According to the result of separation of BTEX by 10-CPM, the R of BTEX not only have above 80% with relative lower pressure but also have higher water flux as compared of other nano-filtration. This suggests that the 10-CPMs possess good potential for BTEX removal in wastewater treatment.

Keywords

CNT Membrane, BTEX, Nanofiltration

1. Introduction

The BTEX, volatile organic compounds, which are benzene (B), toluene (T), ethylben-

zene (E) and *p*-xylene (X) are commonly used as solvents in industrial field. The wastewater containing of BTEX is usually discharged into environment from manufacturing, transportation and purposeful disposal sources. The BTEX are toxic and carcinogenic substances, while the presence of excessive amounts of BTEX in aqueous system may have an adverse impact on water quality or them volatized into ambient atmosphere thus endanger public health. It is clear that sustainable, cost effective and efficient wastewater treatment for BTEX is needed. The development of a cost-effective wastewater treatment process for BTEX removal before releasing into the environment is needed in order to meet the growing demand for cleaner water. Membrane filtration process has been employed for this purpose [1].

Carbon nanotubes (CNT) are unique, one-dimensional macromolecules that possess outstanding gas and water permeability, nanofluidics and molecular transport at nanoscales [2]-[5]. Literatures have demonstrated that the CNT membrane has extremely high transport rates of water and gas due to its unique nano-fluidic system [6]-[14]. These properties lead membrane filtration to high flux and low operation pressure and thus directly benefit the fields of water and wastewater treatment, desalination, gas pollution control, gas/chemical separations and drug delivery [13]-[19]. However, to the knowledge of the authors, filtration of BTEX from aqueous solutions by CNT/ polymer nanocomposite membrane is still limited in the literature.

This article fabricated the CNT/polyvinylidene fluoride (PVDF) nanocomposite membranes (CPMs) to study their physicochemical properties and separation performance of BTEX from aqueous solutions.

2. Materials and Methods

2.1. Preparation of Carbon Nanotubes Membranes

Commercially available multiwalled CNTs (C_{Tube} -100, CNT Co. Ltd., Korea) with outer diameter of 10 - 40 nm and length of 1 - 25 µm were used to make CNT nanocomposite membranes which has 3.16 nm as average pore diameter. The CPMs were prepared by dissolving CNTs into 16 mL of N, N-dimethylformide (DMF, 99.8% purity, Sigma Aldrich Corp., Mo, USA) at 25°C and sonicating for 1 h. The PVDF (Sigma-Aldrich Corp, Mo, USA) was added into the mixture in a CNT/PVDF ratio of 5 - 15 wt% and then was stirred at 400 rpm and heated to 180°C for 24 h to make sure complete dissolution of the polymer. The CNT/PVDF solution was uniformly casted on a glass plate by means of a hand-casting knife with a knife gap set at 300 µm and then subsequently immersed in a deionized water bath to preserve the membrane (wet phase inversion method). The CPMs were trimmed to a circle shape with a diameter of 7 cm. The PVDF membrane (PM) was also prepared using the same procedure with the exception of CNTs addition.

2.2. Pollutants

The employed B, T, E and X were analytical grade with >99% purity and purchased from Merck (Darmstadt, Germany for B and T; Hohenbrunn, Germany for E and X).

These chemical agents were diluted using deionized H_2O to the desired concentrations in enclosed bottles. The molecule diameters of B, T, E and X are respectively 0.58, 0.6, 0.63 and 0.63 nm [20].

2.3. Filtration Experiments

Figure 1 shows the membrane filtration experiment setup. It was conducted by using a crossflow filtration system, which equipped with a pressure vessel containing the membrane module, pressurization tank, a temperature control box and a digital analytical balance. The BTEX containing solution was pressured by compressed air from the tank to the membrane cell within a stainless steel disk, which has inner diameter of 5 cm and an active area of 19.63 cm². The permeation was analyzed for the BTEX concentration and weight to measure the permeation volume. The rejection coefficient of the BTEX concentration, *R*, was calculated as:

$$R = \frac{C_f - C_p}{C_f} \tag{1}$$

where C_f and C_p are the BTEX concentrations in the feed and permeate, respectively. The membrane is completely permeable when R reaches zero while the membrane is completely impermeable when R achieves unity.

2.4. Analytic Methods

The concentrations of BTEX were determined by a gas chromatograph and flame ionization detection (GC-FID, Model Shimadzu 2010, Shimadzu Instruments, CA, USA). The BTEX samples were collected to 5 ml from filter and then injected to purge and trap concentrator (Stratum PTC, Teledyne Tekmar, OH, USA) under trap adsorption temperature at 25°C and trap desorption temperature at 250°C with N₂ as purge gas, then the gaseous BTEX were injected to GC-FID for analysis at detector temperature of 200°C.



Figure 1. Membrane filtration setup.

The CNT content of CPMs was determined by a thermo gravimetric analyzer (TGA I 1000, Instrument Specialists Incorporated, Twin Lakes, WI, USA) at a heating rate of 2°C/min from 25°C - 800°C.

The morphologies of the PM and CPMs were investigated by a field emission scanning electron microscope (FE-SEM, Ultra Plus, Carl Zeiss SMT AG, Germany). Representative samples of the membranes were cut into 1 - 2 mm², attached with carbon tape to aluminum stabs, and shadowed with platinum prior to the SEM measurements.

Membrane surface topography was determined by atomic force microscope (AFM, SPA 400, Seiko Instruments Inc., Japan) image and analysis. Topographic imaging was performed in dynamic force mode with noncontact silicon cantilevers (NSC15/AIBS, spring const. = 20 N/m, Mikromasch, Estonia). The surface roughness of membrane (abbreviated as R_{RMS}) that was typically used to quantify variations in surface elevation was determined directly from the height data by root-mean-square roughness calculation.

The pore properties of the membrane were determined by N_2 adsorption/desorption at 77 K via Micromeritics ASAP 2020 volumetric sorption analyzer (Norcross, GA, USA). The N_2 adsorption/desorption isotherms were measured at a relative pressure range of 0.0001 - 0.99 and then employed to determine specific pore volume and average pore diameter via the micropore (MP) method for pore sizes below 1.7 nm and the Barrett-Johner-Halenda (BJH) method for pore sizes 1.7 - 300 nm.

3. Results and Discussion

3.1. Characterization of the Membrane

3.1.1. TGA

Figure 2 shows the TG curves of PM and 5% - 15% of CNT content CPMs. There are two temperatures for the start of rapid weight loss, named as T_1 and T_2 , which were



Figure 2. TG curves of PM and various CPMs.



ascribed to initial temperatures for thermal decomposition of PVDF and CNTs/ash, respectively [21]. The T_1 of PM, 5-CPM, 10-CPM and 15-CPM were in the range of 362° C - 390° C and T_2 of them was 510° C. The CNTs starts thermal decomposition since T_2 and the weight remained at T_2 can explain to percentage of CNTs content in membrane. The remaining weight of 5, 10, 15-CPMs at T_2 was about 5.1, 9.4 and 14.7 wt%, respectively, which reveals that the real CNTs content in CPMs has well controlled in this research. The residue material at 800°C about 1.0 wt% could be attributed to ash and metallic catalytic from CNTs [22].

3.1.2. Pore Size Distribution

Figure 3 displays the pore size distributions of membranes with micropores (<2 nm), mesopores (2 - 50 nm) and macropores (50 - 300 nm). It can be seen that the PM has narrow micropores at 0.42 nm and has lower amount in mesopores/macropores volume. After adding CNTs into PM, the micropore size range has increased wider to 0.4 - 0.46 nm which can be attributed to the CNTs provided more pores from space of graphite interlayers in the range of 0.3 - 0.4 nm [23] [24]. The mesocopore/macropore volume of PM and CPMs were tenth of magnitude lower than micropore volumes of that. The CNTs content could effects mesocopore/macropore volume of CPM that has increased among of 5 - 70 nm with rising of CNT contents. It can be explained that the CNTs have a number of CNT cavity pores (10 nm) and aggregated pores (40 - 60 nm) which creating the pores in CPMs [25].



Figure 3. Pore size distributions of PM and various CPMs.

3.1.3. FE-SEM

Figures 4(a)-(d) show the FE-SEM images of PM and CPMs surfaces. The PM surface can be seen smoother than CPMs surfaces. While the CNTs added into membranes, the membrane surface became rougher and created few surface macropores. The surface of all CPMs have obviously rugae-like structure that has space below 3 nm can be seen in Figure 4(b-1). The surface macropores were in the range of 10 - 25 nm for 5-CPM, 10 -35 nm for 10-CPMs and 10 - 130 nm for 15-CPMs, respectively, which can prove the CNT content effects on pore size of CPMs. The surface macropore quantity was increased with increasing CNT contents, which can be attributed to the CNTs has aggregated structure and the remained air bubbles in the aggregated pores during CPMs fabrication.

3.1.4. Membrane Surface Roughness (R_{RMS})

R_{RMS} was measured by atomic force microscopy (AFM) which can explain the surface roughness of membrane and its relation of water permeation. The higher value of R_{RMS} reveals the membrane surface is rougher, more hydrophilic, more pore density and more pore volume that are related to filtration performance of membrane [26]. Table 1 shows the R_{RMS} roughness for PM and 5 - 15 CPMs membranes. The R_{RMS} of PM was 35.6 nm which is lower than that of all CPMs. It's the fact that increasing the addition of CNTs in polymer membrane can greater the surface roughness and thus changed the filtration performance.

3.1.5. Optical Contact Angle of Membranes (θ)

Table 2 shows the optical contact angle (θ) of PM and CPMs. The θ of PM was less



Figure 4. SEM images of PM and various CPMs.



Membranes	R _{RMS} (nm)
РМ	35.6
5-CPM	43.5
10-CPM	58.8
15-CPM	78.9

 Table 1. Surface roughness of membranes.

Table 2. Optical contact angle of membranes.

Membrane	$ heta(\degree)$
PM	83.75 ± 2.36
5-CPM	76.07 ± 0.7
10-CPM	71.78 ± 2.03
15-CPM	59.95 ± 1.46

than 90° which reveals the PM has the hydrophilic surface and has good affinity for water [27]. The θ of CPM were decreased with increasing CNT content and R_{RMS}. It is the fact that the addition of CNT into membrane can enhance the hydrophilicity and roughness for membrane thus increasing the water permeation.

3.2. Effect of CNTs Content on Water Permeation

Figure 5 displays the water permeation flux (J_v) of PM and all CPMs with operation time under permeation pressure 10 psi. The J_v of PM and 5-CPM have stable J_v from 10 to 420 min, but that of 10-CPMs and 15-CPMs have decreased then reached equilibrium after 300 min and 420 min, respectively. The decrease of J_v for 10-CPM and 15-CPM are the conventional trend which relating to pore structure of membrane. The large pore size and larger R_{RMS} could cause the water molecular accumulated on membrane and thus reducing the J_v . The similar finding was reported in literatures [28] [29]. The J_v at equilibrium time for PM, 5, 10, 15-CPMs are 2.0, 3.3, 4.6 and 7.1 L/m²·h, respectively. It shows that all CPMs have greater J_v than that of PM, which is a fact that more CNTs content can enhance more water permeability for membrane.

3.3. Separation of BTEX

3.3.1. BTEX Removal

Figure 6 exhibits the *R* of PM and all CPMs at a C_f of 80 mg/L under pressure 10 psi. The *R* of 5-CPM was slightly lower than that of PM for BTEX. The *R* of PM and 5-CPM were respectively 97.4%, 94.3%, 82.6% and 73.8% for B, 99.9%, 97.6%, 87.2% and 79.8% for T, 99%, 97.9%, 89.6% and 85.0% for E, 99.2%, 97.7%, 90.7% and 84.7% for X at the equilibrium time 330 min. It can be seen that the membranes follow the order of PM > 5-CPM > 10-CPM > 15-CPM for BTEX and PM, 5-CPM and 10-CPM have *R* above 80% but 15-CPM. This could be attributed to the increases of mesopoers/macropores from CNTs in 15-CPM resulting some BTEX molecules have penetrated through mem-



Figure 5. Water flux of PM and CPMs at 10 psi permeate pressure.



Figure 6. *R* of BTEX with PM and CPMs.

brane and make higher CNT content membrane have lower BTEX rejection. The 10-CPM is not only has R above 80% for BTEX, but also has greater J_v than 5-CPM. Therefore, 10-CPM has been chosen to exam the following membrane studies.

3.3.2. Effect of Permeation Pressure

Figure 7 shows the *R* and J_v of 10-CPM under various permeation pressure (*P*) from 5 - 40 psi with $C_f = 80$ mg/L of BTEX. The R_j of BTEX has significant decreases with increasing of *P*, but J_v of BTEX has increased with rising of *P*. The *R* of BTEX were more than 90% while *P* was less than 10 psi. Therefore, the *P* at 10 psi was optimal pressure for membrane operation above 90% of *R* for BTEX. The *R* of B, T, E, X were 90.7%, 93.4%, 95.1% and 95.4%, respectively. This shows the order of *R* for BTEX was B < T < E ≈ X that is consistent with the order of molecular weight and molecular size. It indicates that the separation performance of the BTEX from aqueous solutions with CPM is dependent on molecular physical property.

3.3.3. Effect of Feed BTEX Concentrations

Figure 8 displays the effect of C_f in 40, 80 and 120 mg/L with 10-CPM on 10 psi. The *R* of BTEX C_f in 40, 80 and 120 mg/L at 180 min were 87.3%, 80.3% and 68.3% for B, 92.6%, 84.7% and 76.3% for T, 95.0%, 88.5% and 83.9% for E, 96.5%, 90.0% and 85.4%



Figure 7. R and Jv of BTEX for 10-CPMs under various permeation pressures.



Figure 8. Effect of BTEX initial concentration on R with 10-CPMs.

for X, respectively. The result shows that the R of BTEX were decreased with increasing C_f of BTEX. The reason is that the higher C_f can enhance the diffusion force and cause the BTEX molecular to diffuse into membrane. The R of BTEX have also decreased with filtration time, which can be attributed to the concentration accumulated on retentate side with time and then enhanced the diffusion force on membrane, thus reduced R. The R has significant decrease in filtration of B and T as compared with E and X which reveals that the B and T have smaller molecular size and diffused into membrane easier.

3.3.4. Effect of Solution Ionic Strength

The μ , which is a general property of the solution affecting the affinity between the solute and the aqueous solution, was adjusted using NaCl solution. The polar solutes can be affected by μ and adsorbed on membrane surface to become electrical double layer, which could increase the penetration of solutes [30]. **Figure 9** exhibits the effects of solution ionic strength (μ) on the *R* of 10-CPM at a C_f of 80 mg/L and *P* of 10 psi. The *R* of BTEX with μ in 0, 10, 100 mM at 180 min were 80.3%, 80.3% and 72.6% for B, 84.7%, 84.0% and 77.4% for T, 88.5%, 88.2% and 78.1% for E, 89.9%, 87.2% and 80% for X, respectively. It is apparent that the *R* with 10 mM has similar result with that without μ , but *R* with 100 mM was slightly lower (<10%) than that without μ . The result shows that the effect of μ on the *R* was not significant for the μ below 100 mM.

3.3.5. Effect of Solution Temperature

Figure 10 exhibits the effects of solution temperature on the *R* of 10-CPM at a $C_{\rm f}$ of 80 mg/L and *P* of 10 psi. The *R* of BTEX with temperature at 10°C, 20°C, 30°C at 180 min



Figure 9. Effect of μ on *R* of BTEX with 10-CPMs.



Figure 10. Effect of solution temperature on *R* with 10-CPMs.

were 89.4%, 80.3% and 75.0% for B, 92.4%, 84.7% and 79.9% for T, 94.1%, 88.5% and 83.8% for E, 95.2%, 90.0% and 80.9% for X, respectively. It's obviously that the R of BTEX were decreased with increasing temperature, which is due to the thermal expan-

sion of membrane that resultant the pore size enlarger in membrane thus reducing the R of BTEX [31].

3.4. Separation of Mixed BTEX

Figure 11 exhibits the *R* of mixed BTEX at a C_f of 20 and 80 mg/L with 10-CPM at 10 psi. The *R* of mixed B, T, E, X at 180 min were respectively 74.5%, 77.3%, 80.2% and 80.3% for C_f of 20 mg/L and 80.2%, 83.6%, 88.0% and 87.2% for C_f of 80 mg/L. While the C_f is 80 mg/L, the *R* of mixed BTEX is similar to that of individual BTEX (section 3.2.3) and following the same order (B < T < E \approx X) which indicating that the separation mechanism of BTEX can be assigned to size exclusion. The J_v of mixed BTEX at a C_f of 20 and 80 mg/L on 180 min were decreased 21.4% and 34.7%, respectively. It is the fact that the fouling has taken place by concentration on membrane surface and thus decreased the water flux.

3.5. Separation Mechanism

It is observed that the order of R with CPMs was in order of 5-CPM > 10-CPM > 15-CPM is consistent with pore size and specific pore volume. The R of BTEX have effected by permeation pressure, inlet concentration and temperature but not influenced by ionic strength. This indicates that the separation performance of the BTEX from aqueous solutions with CPMs is dependent on size exclusion. The diagram of nanofluid in CPMs and major mechanism for rejecting BTEX from aqueous solutions is graphically presented in Figure 12. The water molecules penetrate through the CPMs with



Figure 11. Filtration of mixed BTEX with CPMs.





Figure 12. Schematic diagram of nanofluid in CPMs and major mechanism for rejecting BTEX from aqueous solutions.

four possible pathways including transport along CNT surface, direct permeation through membrane pores, transport through the cavity pores of CNTs, transport through the aggregated pores and hydrophobic effect [14] [32]. The BTEX were rejected by pore due to size exclusion to reject and thus make the fast separation of BTEX from aqueous solutions. These results reflect that the 10-CPMs possess good potential applications in the separation of the BTEX form aqueous solutions.

4. Conclusion

This article studied the characteristics of PM and CPMs and their separation performance of BTEX from aqueous solutions. The rejection coefficients (R) of BTEX were dependent of the CNT content of CPM, pore size of membrane, molecule size of BTEX, permeation pressure (P), feed concentration (C_i) and temperature, but has no significant influence on ionic strength (μ). The size exclusion plays the important roles on BTEX separation by CPMs. It reveals that the higher molecule size, the higher rejection. The 10-CPM behaved the better separation performance of BTEX with removal efficiencies of 80.3%, 84.7%, 88.5% and 90.0%, respectively, under the relatively lower pressure 10 psi. It has not only higher water flux but also the relatively higher R as comparing of 5% and 15% CNT of CPM. The majority pore size of 10-CPM was <0.4 nm that is capable for removing BTEX which have molecule size < 0.58 nm. This suggests that the 10-CPMs possess good potential for BTEX removal in wastewater treatment.

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