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Effects of Air Pollution on Disease Respiratory: Structures Lag

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Abstract

Objective: Little evidence is available on the association between ozone exposure and health in Campo Grande, Brazil. In this study, we aimed to examine the effects of surface ozone concentrations (03) on respiratory morbidity in Campo Grande, Brazil during the period from January 1st, 2008 to December 31st, 2011. Methods: Daily data on respiratory hospital admissions, O₃, mean temperature, and relative humidity were collected at Campo Grande, Brazil. A Poisson time series model was used to examine the effects of O₃ on hospital admissions, while controlling for seasonality, long-term trend, temperature and relative humidity. A distributed lag non-linear function was used for O₃, temperature, and relative humidity. We examined the effects of O₃ on different age groups (0 - 4 years, 5 - 60 years and >60 years). Results: The ozone-respiratory morbidity relationship was non-linear, with a threshold at 13 ppb (less than 25% percentile of ozone distribution). We estimated the relative risk of hospital admission at 75% percentile of O₃ distribution compared with associated with 25% of percentile of O₃ distribution. The O₃ effect on respiratory morbidity was delayed by two days and lasted for 4 days for all age groups except people aged 5 - 60 years. Children and the elderly were much more vulnerable to ozone pollution than people aged 5 - 60 years. Conclusions: This study suggests that ozone pollution has negative impacts on respiratory diseases in Campo Grande, Brazil. Children and the elderly were susceptible to 03 exposure. These findings should be used to develop policies for protecting people from O₃ pollution.

Kevwords

Air Pollution, Public Health, Respiratory Morbidity, Children, Ozone, Hospital Admission, Elerly

1. Introduction

Atmospheric pollution has been considered as a serious public health issue, especially with the increase of

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population living at urban locations. In Brazil, the population percentage living in cities reached 87% by 2010 and the cities of the developing world will make up 81% of urban humanity [1].

Atmospheric pollution is a hazard factor for human health, even the level of air pollution is very low. The most susceptible people to air pollution are children [2]-[4] and the elderly [5]-[8]. Many studies have reported that there are negative impacts on respiratory mortality and morbidity [2]-[4].

Surface ozone is a pollutant of growing concern in the world [9]. Children are the largest subgroup susceptible to surface ozone [10], particularly in terms of respiratory diseases [11] [12]. However, there are relatively few studies assessing the effects of ozone on children and adolescents' respiratory morbidity.

Moreover, most studies assessing the health effects of ozone levels used relatively short timescales (lags) [13]-[15]. If the effects of ozone on morbidity/mortality were delayed or lasted for a longer period, using short lags would not exactly capture the effects of ozone. The other issue is that most studies use linear function to examine the effects of ozone on morbidity/mortality. The biological manifestations of the effects of pollution on health apparently display behavior that shows a gap between the individual's exposure to pollutants. This means that the attendances/admissions observed on a particular day can be related to pollution of that day, as well as pollution observed in previous days. For a more precise definition of the model to be used, it is essential to determine a lag structure (lag) suitable. We chose to use the model distributed lags up to six days after exposure. Thus, using linear model can not reflect the true ozone-health relationship.

In this study, we aimed to explore the relationship between ozone and the respiratory hospital admissions in Campo Grande, Brazil. We also examined whether the ozone-morbidity relationship was non-linear; whether children and the elderly were more sensitive to ozone air pollution; and whether effects of ozone on morbidity were delayed.

2. Methods

2.1. Data Collection

This study was conducted in Campo Grande, Brazil. Campo Grande city (20°27'16"S; 54°47'16"W, altitude 650 m) is sited on the Maracaju-Campo Grande plateau. The region is tropical with high temperatures predominating throughout the year. It has high rainfall and high relative humidity (Sant' Anna Neto *et al.*, 2003).

We obtained data on health centers' records of out-patient visits from the Secretary for Municipal Health from January 2008 to 31 December 2011. Respiratory diseases were coded using the 9th Revised International Classification of Diseases (ICD-9: 460 to 519). We also stratified the respiratory morbidity by age (0 - 4, 5 - 60, and >60 years).

We obtained the daily mean ozone concentration (O₃ in ppb) from 2008 to 2011. The ozone pollution was measured at the monitoring station of the Physics Department of the Federal University of Mato Grosso do Sul. Ozone concentrations were recorded every 15 minutes, which were used to calculate daily mean concentrations.

We collected daily data on mean temperature, relative humidity from the EMPRAPA research station for beef production in Campo Grande.

2.2. Data Analysis

A quasi-Poisson regression model with distributed lag non-linear model (DLNM) was used to examine the effects of ozone on morbidity. The quasi-Poisson function has the ability to adjust inference for overdispersion [16].

The DLNM allows nonlinear function for exposure and lag to be modeled simultaneously in quite flexible ways [17] [18]. To examine the non-linear ozone-morbidity relationship, a DLNM was used for ozone with 5 degrees of freedom natural cubic spline, and 4 degrees of freedom natural cubic spline was used for lag up to 6 days. We controlled for temperature and relative humidity using a DLNM with 5 degrees of freedom natural cubic spline for exposure (mean temperature and relative humidity) and 4 degrees of freedom natural cubic spline for lag up to 10 days. We controlled for day of the week using category variable. We controlled for seasonality and long-term trend using natural cubic spline with 7 degrees of freedom per year for time.

3. Results

There were 26,941 respiratory hospital admissions during the study period. The daily average concentrations of

O₃ were below the quality standards of ozone pollution (80 ppb) (**Table 1**). The average counts of daily admissions for respiratory diseases were: total admissions for respiratory diseases (11), 0 - 4 years of (5), 5 - 60 years (3), and >60 years (3) (**Table 1**).

Figure 1 shows the time series of respiratory hospital admission, ozone and mean temperature. O₃ had a seasonal trend, with higher concentration in summer than winter. Respiratory hospital admissions were higher in winter than summer.

Figure 2 shows the relationship between ozone and respiratory hospital admission at lag 0 - 6 days. We found that the effects of ozone on respiratory hospital admission were non-linear, with a threshold at 13 ppb. The effects of ozone on children and the elderly were high than people aged 5 - 60 years.

Table 2 shows the estimated relative risks (RR) corresponds to age group: all year risk with interval RR = 1.1079; CI: (1.0379 - 1.1826); risk of 0 - 4 years with a range of RR = 1.1296; CI: (1.0524 - 1.2125), 5 - 60 years, risk with interval of RR=1.014; CI: (0.9410 - 1.0928) and >60 years with risk interval and 1.1546 (1.0406 and 1.2811).

Table 2 and **Figure 3** show the relative risks of hospital admissions at 75% percentile of ozone distribution compared with 25% percentile along the lag days. Results show that the effects on hospitalizations were delayed by two days, and lasted for four days.

4. Discussion

The purpose of this study was to examine the effects of ozone on respiratory hospital admissions in Campo Grande, Brazil during 2008-2011. We found that the ozone-respiratory morbidity relationship was non-linear. The effects of ozone on respiratory morbidity were delayed by two days and lasted for 4 days for all age groups except people aged 5 - 60 years. Children and the elderly were much more vulnerable to ozone pollution than people aged 5 - 60 years.

The magnitude of the association ozone pollution and hospital admissions for respiratory disease was similar to studies conducted in San Paulo [2]-[4]. This might be caused by that the source of zone pollutant in the city is almost the same: the mobile sources, mainly, and stationary sources (industries) and biomass burning in Campo

Table 1. Descriptive analysis of respiratory hospital admission, ozone, and weather conditions in Campo Grande, Brazil during 2008-2011.

	Average	Standard deviation	Minimum	10 quartil	Median	30 quartil	Maximum
Total	11	3	2.00	8.00	10.00	12.00	23.00
0 - 4 years	5	2	1.00	4.00	5.00	6.00	13.00
5 - 60 years	3	1	0.00	2.00	2.00	3.00	8.00
>60 years	3	1	0.00	2.00	3.00	3.00	7.00
Ozono (ppb)	17.56	8	0.66	12.97	16.44	20.73	52.76
Mean Temperature (°C)	23.50	4	7.00	22.10	24.20	25.73	30.81
Relative humidity (%)	66.06	16.12	19.16	55.00	67. 81	78. 83	98.00

Table 2. Relative risk (RR) of respiratory hospital admissions at at 75% percentile of ozone distribution compared with 25% percentile along the lag days in Campo Grande, Brazil during 2008-2011.

	All years	0 - 4 years	5 - 60 years	>60 years	
	RR-CI	RR-CI	RR-CI	RR-CI	
lag 0	1.0035; (0.9758 - 1.0321)	1.0153; (0.9850 - 1.0466)	0.9775; (0.9465 - 1.0095)	1.0028; (0.9587 - 1.0488)	
lag 1	1.0128; (0.9989 - 1.0270)	1.0191; (1.0039 - 1.0345)	0.9921; (0.9764 - 1.0080)	1.0188; (0.9965 - 1.0416)	
lag 2	1.0119; (1.0046 - 1.0336)	1.0211; (1.0056 - 1.0369)	1.0031; (0.9868 - 1.0196)	1.0292; (1.0061 - 1.0530)	
lag 3	1.0209; (1.0056 - 1.0365)	1.0209; (1.0043 - 1.0376)	1.0092; (0.9918 - 1.0269)	1.0320; (1.0073 - 1.0573)	
lag 4	1.0196; (1.0068 - 1.0325)	1.0189; (1.0051 - 1.0328)	1.0115; (0.9970 - 1.0263)	1.0287; (1.0082 - 1.0496)	
lag 5	1.0160; (1.0021 - 1.0302)	1.0157; (1.0006 - 1.0310)	1.0113; (0.9954 - 1.0274)	1.0215; (0.9992 - 1.0444)	
lag 6	1.0114; (0.9883 - 1.0352)	1.0120; (0.9870 - 1.0376)	1.0097; (0.9832 - 1.0369)	1.0125; (0.9754 - 1.0509)	

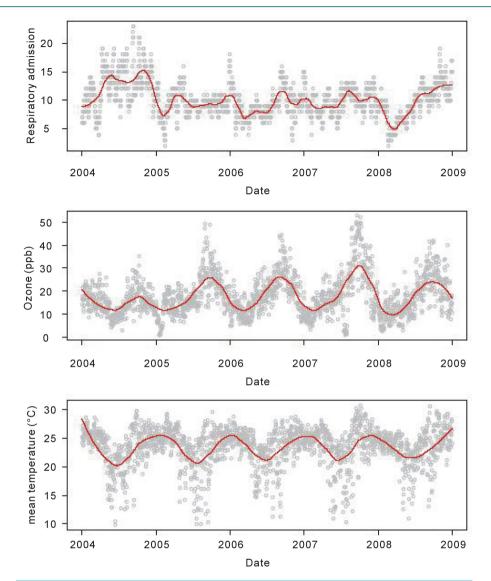


Figure 1. Time series of respiratory hospital admissions, ozone concentration (ppb) and mean temperature (°C) during 2008-2011.

Grande.

Ozone-related epidemiological studies generally assume a linear effect of ozone [19]. However, Lefohn *et al.* reported a nonlinear relationship between ozone dose and pulmonary function [20]. The lag effect of air pollutants is also considered a critical factor in health assessments. This study applied DLNM to calculate the nonlinear association and cumulative risks across lag days for air pollutants. The findings enable greater flexibility when presenting the nonlinear ozone exposure-response curve for outpatient visits for total respiratory diseases (RD).

We found that children and the elderly were much more sensitive to ozone pollution than adult. Children are highly susceptible to exposure to air pollutants. Minute ventilation is higher in children than that in adults because children have higher basal metabolic rates and engage in more physical activity than adults do, as well as because children spend more time outdoors than adults do. On the basis of body weight, the volume of air passing through the airways of a child at rest is twice that of an adult under similar conditions. Pollutant-induced irritation producing a weak response in adults can result in significant obstruction in children. In addition, the fact that their immune system is not fully developed increases the possibility of respiratory infections [21] [22].

Elderly individuals are susceptible to the adverse effects of exposure to air pollutants because they have a less

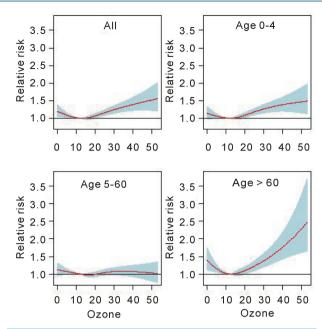


Figure 2. Relationship between ozone and respiratory hospital admission in Campo Grande, Brazil during 2008-2011.

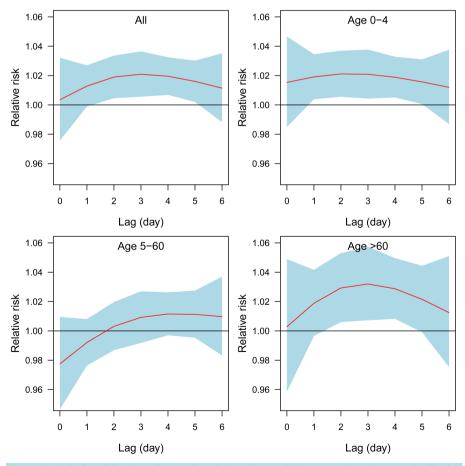


Figure 3. Relative risk of respiratory hospitalizations at 75% percentile of ozone distribution compared with 25% percentile along the lag days in Campo Grande, Brazil during 2008-2011.

efficient immune system (immunosenescence) and a progressive decline in pulmonary function that can lead to airway obstruction and exercise limitation. There is decreased chest wall compliance and lung hyperinflation requiring additional energy expenditure to perform respiratory movements, as well as functional decline of organ systems [23] [24].

Several mechanisms have been suggested to explain the adverse effects of air pollutants. The most consistent and most widely accepted explanation is that, once in contact with the respiratory epithelium, high concentrations of oxidants and pro-oxidants in environmental pollutants such as PM of various sizes and compositions and in gases such as O₃ and nitrogen oxides cause the formation of oxygen and nitrogen free radicals, which in turn induce oxidative stress in the airways. In other words, an increase in free radicals that are not neutralized by antioxidant defenses initiates an inflammatory response with release of inflammatory cells and mediators (cytokines, chemokines, and adhesion molecules) that reach the systemic circulation, leading to subclinical inflammation, which not only has a negative effect on the respiratory system but also causes systemic effects [21].

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